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Chilkoot Lake Sockeye Salmon Stock Status and Escapement Goal Review

by

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Month 2017

Alaska Department of Fish and Game Divisions of Sport Fish and Commercial Fisheries

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**Weights and measures (metric)**

centimeter cm

deciliter dL

gram g

hectare ha

kilogram kg

kilometer km

liter L

meter m

milliliter mL

millimeter mm

**Weights and measures (English)**

cubic feet per second ft3/s

foot ft

gallon gal

inch in

mile mi

nautical mile nmi

ounce oz

pound lb

quart qt

yard yd

**Time and temperature**

day d

degrees Celsius °C

degrees Fahrenheit °F

degrees kelvin K

hour h

minute min

second s

**Physics and chemistry**

all atomic symbols

alternating current AC

ampere A

calorie cal

direct current DC

hertz Hz

horsepower hp

hydrogen ion activity pH

(negative log of)

parts per million ppm

parts per thousand ppt,

‰

volts V

watts W

**General**

Alaska Administrative

Code AAC

all commonly accepted

abbreviations e.g., Mr., Mrs., AM, PM, etc.

all commonly accepted

professional titles e.g., Dr., Ph.D.,

R.N., etc.

at @

compass directions:

east E

north N

south S

west W

copyright ©

corporate suffixes:

Company Co.

Corporation Corp.

Incorporated Inc.

Limited Ltd.

District of Columbia D.C.

et alii (and others) et al.

et cetera (and so forth) etc.

exempli gratia

(for example) e.g.

Federal Information

Code FIC

id est (that is) i.e.

latitude or longitude lat. or long.

monetary symbols

(U.S.) $, ¢

months (tables and

figures): first three

letters Jan,...,Dec

registered trademark ®

trademark ™

United States

(adjective) U.S.

United States of

America (noun) USA

U.S.C. United States Code

U.S. state use two-letter abbreviations (e.g., AK, WA)

**Mathematics, statistics**

*all standard mathematical*

*signs, symbols and*

*abbreviations*

alternate hypothesis HA

base of natural logarithm *e*

catch per unit effort CPUE

coefficient of variation CV

common test statistics (F, t, χ2, etc.)

confidence interval CI

correlation coefficient

(multiple) R

correlation coefficient

(simple) r

covariance cov

degree (angular ) °

degrees of freedom df

expected value *E*

greater than >

greater than or equal to ≥

harvest per unit effort HPUE

less than <

less than or equal to ≤

logarithm (natural) ln

logarithm (base 10) log

logarithm (specify base) log2, etc.

minute (angular) '

not significant NS

null hypothesis HO

percent %

probability P

probability of a type I error

(rejection of the null

hypothesis when true) α

probability of a type II error

(acceptance of the null

hypothesis when false) β

second (angular) "

standard deviation SD

standard error SE

variance

population Var

sample var

fishery manuscript series no. 17-XX

CHILKooT LAKE Sockeye Salmon Stock Status and Escapement Goal Review

By

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Month 2017

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# abstract

Chilkoot Lake, located in upper Lynn Canal near the city of Haines, supports one of the largest runs of sockeye salmon (*Oncorhynchus nerka*) in southeast Alaska. This stock is currently managed as a sustainable escapement goal (SEG) range with a lower bound of 38,000 and an upper bound of 86,000 spawners. Escapement is monitored by ADF&G with a weir on the Chilkoot River and stock of origin from the District 15 commercial harvest is determined using scale pattern analysis. We used Ricker spawner-recuit models in a Bayesian framework to fit data from brood years 1976–2010. Given significant autocorrelation at lag-1, we chose an autoregressive Ricker model (AR1) for this assessment. Based on model results, maximum sustainable yield would be achieved with an escapement of approximately 52,900 sockeye salmon (median of Smsy) and a range of 45,000–60,000 spawners would result in a greater than 80% probability of achieving at least 90% of maximum sustainable yield. This range of escapements fits within the current escapement goal range and, given considerable uncertainty in parameter estimates, we do not recommend changes to the goal at this time. However, large escapements since 2012 will provide some contrast once the resulting recruits can be enumerated; thus, we recommend reassessing this escapement goal prior to the Alaska Board of Fisheries meeting in 2021.

Key words: Bayesian statistics, escapement goal, maximum sustained yield, missing data, sockeye salmon, *Oncorhynchus nerka*, Chilkoot Lake, spawner-recruit analysis.

# 

# introduction

The Chilkoot and Chilkat river watersheds, located in northern Southeast Alaska near the town of Haines, support two of the largest sockeye salmon (*Oncorhynchus nerka*) runs in Southeast Alaska (Figure 1). Between 1900 and 1920, the annual commercial harvest of sockeye salmon in northern Southeast Alaska averaged 1.5 million fish, the majority of which were believed to originate from Chilkat and Chilkoot river watersheds (Rich and Ball 1933). Over the past 2 decades, the average sockeye salmon harvest in northern Southeast Alaska was 0.5 million fish, of which an average 96,000 fish originated from Chilkat Lake and 65,000 fish originated from Chilkoot Lake (Eggers et al. 2010). Historically, Chilkoot Lake sockeye salmon were harvested in the large fish trap and purse seine fisheries in Icy and northern Chatham straits as well as in terminal drift gillnet areas of Lynn Canal. Fish traps were eliminated with Alaska statehood in 1959 and Lynn Canal developed into a designated drift gillnet fishing area (District 15) where most of the commercial harvest of Chilkoot Lake sockeye salmon takes place (Figure 1). A smaller portion of the Chilkoot Lake run is harvested in the commercial purse seine fisheries that target pink salmon (*O. gorbuscha*) in Icy and northern Chatham straits. Annual contributions to those fisheries are not known and likely vary annually depending on fishing effort and the strength of pink salmon runs. Chilkoot Lake sockeye salmon are also harvested annually in subsistence fisheries in Chilkoot Inlet and Lutak Inlet, with reported harvests for the period 1990–2016 averaging approximately 2,100 fish per year.

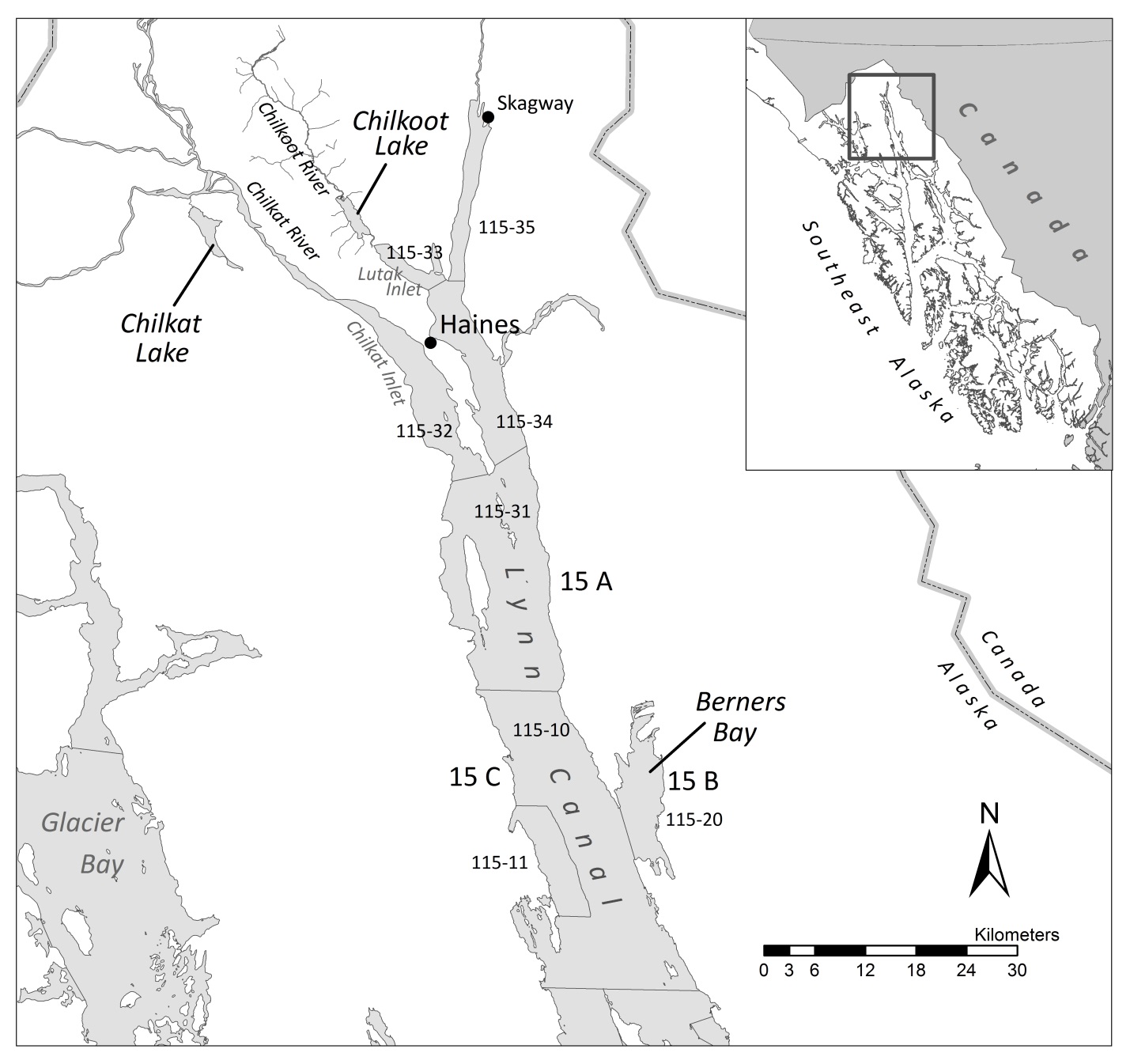


Figure 1.–Commercial fishing subdistrict and management boundary lines within District 15 in the Haines area, Southeast Alaska.

The Alaska Department of Fish and Game (ADF&G) initiated a scale pattern analysis program in 1980 to estimate contributions of sockeye salmon stocks to the District 15 commercial drift gillnet fishery. Bergander (1974) first developed a dichotomous key to classify sockeye salmon scale samples from the fishery as Chilkoot Lake or Chilkat Lake fish, based on distinct differences in their freshwater scale patterns (Stockley 1950). Marshall et al. (1982) improved the sample design and estimated stock contributions using linear discriminant function analysis. McPherson and Marshall (1986) showed that all age classes of the 2 stocks could be identified accurately using a visual classification technique and blind testing procedure. That technique was expanded to include a group of “other” stocks—a combination of Chilkat River mainstem and Berners Bay stocks that contribute to early-season harvests in Lynn Canal (McPherson 1987a). Blind tests to verify accuracy and correct for misclassification have not been conducted since the early 1990s; however, historical stock-specific harvest estimates based solely on visual classification were highly accurate and the difference between initial and corrected estimates varied by only 2% or less (McPherson and Marshall 1986; McPherson 1987a, 1987b; McPherson and Jones 1987; McPherson 1989; McPherson et al. 1992; McPherson and Olsen 1992). The consistent differences in freshwater scale patterns makes visual scale pattern analysis highly accurate, and it is more cost effective and requires less time than other stock-identification methods (McPherson 1990; McPherson and Olsen 1992).

Chilkoot Lake sockeye salmon escapements have been counted annually through an adult counting weir on the Chilkoot River since 1976 (Bachman and Sogge 2006; Bachman et al. 2013 and 2014). The run has 2 components, an early and a late run, which were managed as separate units through 2005 (Geiger et al. 2005). Total annual weir counts averaged 80,000 sockeye salmon through 1993, but declined to an average of only 30,000 fish from 1994 to 2000. Weir counts have averaged 68,000 fish since 2000. In addition to salmon counts, biological data have been collected annually at the weir to estimate age, size, and sex composition of the escapement and for use in scale pattern analysis. Basic information about lake productivity and rearing sockeye salmon fry populations has also been collected through limnological and hydroacoustic sampling conducted most years since 1987 (Barto 1996; Riffe 2006; Bachman et al. 2014). Those studies have been used to assess potential sockeye salmon production from the lake (Barto 1996).

The Chilkoot Lake run has been managed for at least 5 different escapement goals since 1976. Informal goals of 80,000–100,000 fish (1976–1980) and 60,000–80,000 fish (1981–1989; Bergander et al. 1988) were replaced in 1990 by a biological escapement goal of 50,500–91,500 sockeye salmon (McPherson 1990). The goal was divided into separate goals for early (16,500– 31,500 fish) and late runs (34,000–60,000 fish). In 2006, the escapement goal was rounded to 50,000–90,000 sockeye salmon and classified as a sustainable escapement goal due to uncertainty in escapement levels based on weir counts (Geiger et al. 2005). Early- and late-run goals were eliminated and replaced with weekly cumulative escapement targets based on historical run timing. The existing sustainable escapement goal of 38,000–86,000 sockeye salmon was established in 2009 based on an autoregressive Ricker spawner-recruit model by Eggers et al. (2009) that relied on brood year escapement and returns data from 1976–2003. Specifically, the recommended escapement goal by Eggers et al. (2009) was the range of spawners expected to produce at least 90% of MSY. The goals of this study are to update the analysis by Eggers et al. in a Bayesian framework and included escapement and return data from brood years 1976-2009.

# 

# STUDY SITE

Chilkoot Lake (ADF&G Anadromous Waters Catalogue No. 115-33-10200-0010; 59°21′16” N, 135°35′42” W) is located at the head of Lutak Inlet, approximately 16 km northeast of the city of Haines, Alaska (Figures 1 and 2). It is glacially turbid, has a surface area of 7.2 km2 (1,734 acres), a mean depth of 55 m, a maximum depth of 89 m, and a total volume of 382.4 × 106 m3. The Chilkoot River begins at glacier terminuses east of the Takshunak Mountains and west of the Ferebee Glacier. The glacial river flows approximately 26 km southeast into Chilkoot Lake, then flows approximately 2 km into Lutak Inlet. Early-run sockeye salmon spawn in small lake and river tributaries and late-run fish spawn in the main channel of the Chilkoot River and along lake beaches where upwelling water occurs (McPherson 1990). Chilkoot Lake is located within the northern temperate rainforest that dominates the Pacific Northwest coast of North America. Although the climate is characterized by cold winters and cool, wet summers, the lake is set in a transitional zone, with warmer and drier summers and cooler winters than the rest of Southeast Alaska (Bieniek et al. 2012). Average precipitation in the study area is approximately 165 cm/year (Bugliosi 1988). Sitka spruce (*Picea sitchensis*), western hemlock (*Tsuga heterophylla*), and Sitka alder (*Alnus viridis*) dominate the forested watershed.



Figure .–Map showing Lutak Inlet, Chilkoot Lake, and the location of the limnology stations and salmon counting weir.

# Methods

## Data

### Harvest Estimates

Commercial harvest data for the District 15 commercial drift gillnet fishery in northern Lynn Canal was obtained from the ADF&G Southeast Alaska Integrated Fisheries Database. However, harvest from District 15 contains sockeye salmon from multiple stocks. Thus, visual scale pattern analysis (SPA) was used to determine stock composition of sockeye salmon harvested in the District 15 commercial drift gillnet fishery and estimate harvest of fish bound for Chilkoot Lake (Bachman et al. 2014). The general methods of stock apportionment using SPA have remained unchanged since the mid-1980s: escapement scale samples from 3 stocks of known origin, Chilkoot Lake, Chilkat Lake, and “other” (Chilkat River mainstem and Berners Bay stocks), were aged and compared to scale samples from the commercial fisheries, which were apportioned to these three stocks for each statistical week. Since total District 15 commercial drift gillnet harvest was not apportioned to a particular stock in years 1976 through 1983, the apportionment percentages from McPherson (1990) were reapplied to updated harvest from those years (Bednarski et al. 2017).

Table 1.–Estimated total harvest of Chilkoot Lake sockeye salmon and harvest proportions by age, 1976–2016.

|  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- |
|  |  | **Harvest proportion by age** | | |  |  |  |  |
| **Year** | **Harvest** | **Age 2** | **Age 3** | **Age 4** | **Age 5** | **Age 6** | **Age 7** | **Age 8** |
| 1976a | 58,765 | 0.00 | 0.00 | 0.05 | 0.70 | 0.25 | 0.00 | 0.00 |
| 1977a | 41,477 | 0.00 | 0.00 | 0.03 | 0.71 | 0.26 | 0.00 | 0.00 |
| 1978a | 89,558 | 0.00 | 0.00 | 0.04 | 0.63 | 0.33 | 0.00 | 0.00 |
| 1979a | 115,995 | 0.00 | 0.00 | 0.03 | 0.84 | 0.13 | 0.01 | 0.00 |
| 1980a | 31,267 | 0.00 | 0.00 | 0.02 | 0.53 | 0.44 | 0.00 | 0.00 |
| 1981a | 48,420 | 0.00 | 0.00 | 0.03 | 0.47 | 0.50 | 0.01 | 0.00 |
| 1982a | 127,174 | 0.00 | 0.00 | 0.01 | 0.44 | 0.54 | 0.00 | 0.00 |
| 1983a | 124,180 | 0.00 | 0.00 | 0.02 | 0.45 | 0.53 | 0.00 | 0.00 |
| 1984 | 99,592 | 0.00 | 0.00 | 0.00 | 0.70 | 0.30 | 0.00 | 0.00 |
| 1985 | 131,091 | 0.00 | 0.00 | 0.00 | 0.33 | 0.66 | 0.00 | 0.00 |
| 1986 | 168,006 | 0.00 | 0.00 | 0.01 | 0.37 | 0.61 | 0.00 | 0.00 |
| 1987 | 69,900 | 0.00 | 0.00 | 0.01 | 0.60 | 0.39 | 0.00 | 0.00 |
| 1988 | 76,883 | 0.00 | 0.00 | 0.03 | 0.46 | 0.51 | 0.00 | 0.00 |
| 1989 | 156,160 | 0.00 | 0.00 | 0.00 | 0.68 | 0.32 | 0.00 | 0.00 |
| 1990 | 149,377 | 0.00 | 0.00 | 0.01 | 0.45 | 0.54 | 0.00 | 0.00 |
| 1991 | 60,721 | 0.00 | 0.00 | 0.01 | 0.39 | 0.60 | 0.01 | 0.00 |
| 1992 | 113,146 | 0.00 | 0.00 | 0.01 | 0.46 | 0.53 | 0.00 | 0.00 |
| 1993 | 103,531 | 0.00 | 0.00 | 0.02 | 0.32 | 0.66 | 0.00 | 0.00 |
| 1994 | 126,852 | 0.00 | 0.00 | 0.01 | 0.55 | 0.41 | 0.02 | 0.00 |
| 1995 | 68,737 | 0.00 | 0.00 | 0.04 | 0.46 | 0.50 | 0.00 | 0.00 |
| 1996 | 99,677 | 0.00 | 0.00 | 0.02 | 0.48 | 0.50 | 0.00 | 0.00 |
| 1997 | 73,761 | 0.00 | 0.00 | 0.04 | 0.39 | 0.57 | 0.00 | 0.00 |
| 1998 | 112,630 | 0.00 | 0.00 | 0.02 | 0.68 | 0.30 | 0.00 | 0.00 |
| 1999 | 149,410 | 0.00 | 0.00 | 0.01 | 0.46 | 0.52 | 0.00 | 0.00 |
| 2000 | 78,265 | 0.00 | 0.00 | 0.02 | 0.12 | 0.85 | 0.00 | 0.00 |
| 2001 | 60,183 | 0.00 | 0.00 | 0.03 | 0.76 | 0.19 | 0.02 | 0.00 |
| 2002 | 47,332 | 0.00 | 0.00 | 0.03 | 0.45 | 0.52 | 0.00 | 0.00 |
| 2003 | 49,955 | 0.00 | 0.00 | 0.02 | 0.28 | 0.70 | 0.00 | 0.00 |
| 2004 | 51,110 | 0.00 | 0.00 | 0.05 | 0.69 | 0.25 | 0.01 | 0.00 |
| 2005 | 22,852 | 0.00 | 0.00 | 0.03 | 0.36 | 0.62 | 0.00 | 0.00 |
| 2006 | 15,979 | 0.00 | 0.00 | 0.08 | 0.53 | 0.38 | 0.01 | 0.00 |
| 2007 | 14,208 | 0.00 | 0.00 | 0.02 | 0.64 | 0.33 | 0.01 | 0.00 |
| 2008 | 22,156 | 0.00 | 0.00 | 0.04 | 0.82 | 0.14 | 0.00 | 0.00 |
| 2009 | 85,551 | 0.00 | 0.00 | 0.01 | 0.51 | 0.48 | 0.00 | 0.00 |
| 2010 | 48,079 | 0.00 | 0.00 | 0.01 | 0.24 | 0.74 | 0.00 | 0.00 |
| 2011 | 15,599 | 0.00 | 0.00 | 0.03 | 0.62 | 0.35 | 0.00 | 0.00 |
| 2012 | 54,884 | 0.00 | 0.00 | 0.08 | 0.69 | 0.22 | 0.00 | 0.00 |
| 2013 | 75,588 | 0.00 | 0.00 | 0.03 | 0.53 | 0.44 | 0.00 | 0.00 |
| 2014 | 81,502 | 0.00 | 0.00 | 0.04 | 0.64 | 0.31 | 0.01 | 0.00 |
| 2015 | 33,085 | 0.00 | 0.00 | 0.02 | 0.32 | 0.66 | 0.00 | 0.00 |
| 2016 | 35,991 | 0.00 | 0.00 | 0.05 | 0.55 | 0.40 | 0.01 | 0.00 |

a. McPherson 1990.

### Escapement Estimates

Sockeye salmon entering into Chilkoot Lake have been counted through a weir on the Chilkoot River, located downstream of the lake outlet, from 1976 through 2017 (Bergander 1989, 1990; Bachman and Kelley 1999; Bachman 2003; Bachman and Sogge 2006; Bednarski et al. 2017). The run has two components, an early and a late run, and these two components are currently managed as a single unit. The sockeye salmon weir counts have varied dramatically during these years, from 7,200 (1995) to 103,000 (1982) fish (Table 1, Figure 3). Weir counts have averaged 66,273 sockeye salmon between 1976 and 2016. Weir counts were low during the period, 1994 to 2000, and averaged a little over 30 thousand during this period. Escapement age compositions are based on annual scale samples taken at the weir and used to estimate escapement by age (Table 2).

The extremely low weir count in 1995 prompted ADF&G to verify the weir counts by conducting mark-recapture experiments on Chilkoot Lake sockeye salmon. The mark-recapture project was conducted annually from 1996 to 2004 and again in 2007, (Bachman and Kelley 1999, Bachman and Sogge 2006, Bachman and Eisenman *in prep*.). The mark recapture estimates were consistently higher than weir counts averaging 1.84 times the weir count (Table 1). Because spawning in Chilkoot Lake occurs primarily in beach spawning areas and in the remote upper reaches of the Chilkoot watershed, the second-event recovery is difficult and low tag recoveries have contributed to imprecise mark recapture estimates. Differences between mark recapture were not consistent enough for a calibration of the weir counts (Figure 4). Assessments of Chilkoot sockeye salmon escapements are based on weir counts, recognizing that they are likely conservative.

### Recruits from Parent Escapement by Age

Scale samples from commercial harvests and escapement were analyzed at the ADF&G salmon-aging laboratory in Douglas, Alaska. Age classes were designated by the European aging system where freshwater and saltwater years were separated by a period (e.g., 1.3 denoted a fish with 1 freshwater and 3 ocean years; Koo 1962). The weekly age distribution (the seasonal age distribution weighted by week) calculated using equations from Cochran (1977).

The recruits, by age, from parent escapements were estimated for the 1976 to 2010 brood years (Table 4). The recruits from brood year *y* and age *a* is the escapement and catch for age *a* in calendar year *y + a*.

 (1)

*Ra,y* is the recruits for age a and brood year *y, Ea,y+a* is the escapement by age a and calendar year *y+a*, and *Ca,y+a* is catch by age *a* and calendar year *y+a*.

Production for year classes 1976 through 2010 was estimated for each cohort as the sum of production at age over ages of the cohort:

 (2)

As of this writing, some of the older and rarer age class from the 2010 brood year had not yet returned and been enumerated. However, based on previous years, the incomplete age classes were estimated to represent less than 1% of the total brood year recruits and therefore we consider it unlikely that these will have a substantial influence on the results of our analysis.

Table 2.–Expanded Chilkoot Lake sockeye salmon escapement estimates (1976-2016); mark–recapture (MR)

|  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
|  |  |  |  | **Escapement proportion by age** | | | | | | |  | |  | |  | |  | |
| **Year** | **DIDSON** | **MR** | **Weir** | **Age 2** | | **Age 3** | | **Age 4** | | **Age 5** | | | **Age 6** | | **Age 7** | | **Age 8** | |
| 1976a | NA | NA | 69,729 | 0.00 | 0.00 | | 0.06 | | 0.65 | | | 0.29 | | 0.00 | | 0.00 | |
| 1977a | NA | NA | 50,363 | 0.00 | 0.00 | | 0.04 | | 0.77 | | | 0.19 | | 0.00 | | 0.00 | |
| 1978a | NA | NA | 67,528 | 0.00 | 0.00 | | 0.10 | | 0.66 | | | 0.25 | | 0.00 | | 0.00 | |
| 1979a | NA | NA | 80,588 | 0.00 | 0.00 | | 0.01 | | 0.71 | | | 0.28 | | 0.00 | | 0.00 | |
| 1980a | NA | NA | 101,135 | 0.00 | 0.00 | | 0.02 | | 0.64 | | | 0.33 | | 0.00 | | 0.00 | |
| 1981a | NA | NA | 84,097 | 0.00 | 0.00 | | 0.00 | | 0.59 | | | 0.41 | | 0.00 | | 0.00 | |
| 1982 | NA | NA | 86,213 | 0.00 | 0.01 | | 0.04 | | 0.60 | | | 0.36 | | 0.00 | | 0.00 | |
| 1983 | NA | NA | 134,601 | 0.00 | 0.01 | | 0.06 | | 0.65 | | | 0.28 | | 0.00 | | 0.00 | |
| 1984 | NA | NA | 123,190 | 0.00 | 0.00 | | 0.03 | | 0.76 | | | 0.21 | | 0.00 | | 0.00 | |
| 1985 | NA | NA | 58,335 | 0.00 | 0.01 | | 0.04 | | 0.49 | | | 0.46 | | 0.00 | | 0.00 | |
| 1986 | NA | NA | 23,947 | 0.00 | 0.00 | | 0.05 | | 0.29 | | | 0.66 | | 0.00 | | 0.00 | |
| 1987 | NA | NA | 48,972 | 0.00 | 0.01 | | 0.05 | | 0.59 | | | 0.34 | | 0.00 | | 0.00 | |
| 1988 | NA | NA | 27,722 | 0.00 | 0.00 | | 0.01 | | 0.38 | | | 0.61 | | 0.00 | | 0.00 | |
| 1989 | NA | NA | 141,475 | 0.00 | 0.00 | | 0.01 | | 0.71 | | | 0.28 | | 0.00 | | 0.00 | |
| 1990 | NA | NA | 60,230 | 0.00 | 0.00 | | 0.02 | | 0.42 | | | 0.56 | | 0.00 | | 0.00 | |
| 1991 | NA | NA | 51,138 | 0.00 | 0.00 | | 0.02 | | 0.55 | | | 0.42 | | 0.00 | | 0.00 | |
| 1992 | NA | NA | 95,880 | 0.00 | 0.00 | | 0.01 | | 0.42 | | | 0.57 | | 0.00 | | 0.00 | |
| 1993 | NA | NA | 212,757 | 0.00 | 0.00 | | 0.04 | | 0.41 | | | 0.55 | | 0.00 | | 0.00 | |
| 1994 | NA | 153,540 | 86,385 | 0.00 | 0.00 | | 0.02 | | 0.60 | | | 0.37 | | 0.01 | | 0.00 | |
| 1995 | NA | 184,541 | 61,783 | 0.00 | 0.00 | | 0.04 | | 0.46 | | | 0.49 | | 0.00 | | 0.00 | |
| 1996b | NA | 262,852 | NA | NA | NA | | NA | | NA | | | NA | | NA | | NA | |
| 1997b | NA | 238,803 | NA | NA | NA | | NA | | NA | | | NA | | NA | | NA | |
| 1998b | NA | 211,114 | NA | NA | NA | | NA | | NA | | | NA | | NA | | NA | |
| 1999 | NA | 240,002 | 134,048 | 0.00 | 0.00 | | 0.01 | | 0.40 | | | 0.59 | | 0.00 | | 0.00 | |
| 2000 | NA | 132,687 | 47,077 | 0.00 | 0.00 | | 0.02 | | 0.14 | | | 0.84 | | 0.00 | | 0.00 | |
| 2001 | NA | 105,064 | 53,239 | 0.00 | 0.00 | | 0.03 | | 0.79 | | | 0.16 | | 0.02 | | 0.00 | |
| 2002 | NA | 148,465 | 65,611 | 0.00 | 0.00 | | 0.03 | | 0.45 | | | 0.52 | | 0.00 | | 0.00 | |
| 2003 | NA | 116,891 | 55,516 | 0.00 | 0.00 | | 0.03 | | 0.30 | | | 0.67 | | 0.00 | | 0.00 | |
| 2004 | NA | 118,795 | 83,534 | 0.00 | 0.00 | | 0.04 | | 0.65 | | | 0.30 | | 0.01 | | 0.00 | |
| 2005 | NA | 89,072 | 32,098 | 0.00 | 0.00 | | 0.04 | | 0.39 | | | 0.57 | | 0.00 | | 0.00 | |
| 2006 | NA | 91,439 | 38,850 | 0.00 | 0.00 | | 0.05 | | 0.62 | | | 0.32 | | 0.00 | | 0.00 | |
| 2007 | NA | 59,884 | 27,915 | 0.00 | 0.00 | | 0.07 | | 0.48 | | | 0.45 | | 0.00 | | 0.00 | |
| 2008 | 74,919 | 119,808 | NA | 0.00 | 0.00 | | 0.05 | | 0.82 | | | 0.12 | | 0.00 | | 0.00 | |
| 2009 | 153,033 | 285,218 | NA | 0.00 | 0.00 | | 0.03 | | 0.71 | | | 0.25 | | 0.00 | | 0.00 | |
| 2010 | 61,906 | 72,318 | NA | 0.00 | 0.01 | | 0.03 | | 0.37 | | | 0.59 | | 0.00 | | 0.00 | |
| 2011 | 63,628 | 109,335 | NA | 0.00 | 0.00 | | 0.09 | | 0.61 | | | 0.29 | | 0.00 | | 0.00 | |
| 2012 | 121,810 | 171,924 | NA | 0.00 | 0.01 | | 0.05 | | 0.77 | | | 0.17 | | 0.00 | | 0.00 | |
| 2013 | 116,300 | 224,516 | NA | 0.00 | 0.00 | | 0.08 | | 0.49 | | | 0.43 | | 0.00 | | 0.00 | |
| 2014 | 70,470 | 212,201 | NA | 0.00 | 0.01 | | 0.02 | | 0.72 | | | 0.25 | | 0.01 | | 0.00 | |
| 2015 | 175,874 | 124,892 | NA | 0.00 | 0.00 | | 0.04 | | 0.29 | | | 0.66 | | 0.00 | | 0.00 | |
| 2016 | 88,513 | 96,148 | NA | 0.00 | 0.00 | | 0.04 | | 0.60 | | | 0.36 | | 0.00 | | 0.00 | |

a.The weir was not operated in 1996, 1997, or 1998.

Table 3.–Number of enhanced sockeye salmon fry in Chilkoot Lake for release years 1989 to 2003 (Eggers et al. 2010). The stocked fry were incubated as eggs in the hatchery and released as fry into Chilkoot Lake in the spring. The number of fry that emerged in the spring each year from incubation boxes was estimated from the number of eggs seeded in incubation boxes minus the dead eggs counted in the spring.

|  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- |
| **Brood Year** | **Release Year** | **Stocked fry (thermal marked)a** | **Stocked fry (unmarked)** | **Incubation fry boxb** | **Total enhanced fry** |
| 1988 | 1989 | 0 | 0 | 15,094 | 15,094 |
| 1989 | 1990 | 0 | 0 | 300,127 | 300,127 |
| 1990 | 1991 | 0 | 0 | 388,000 | 388,000 |
| 1991 | 1992 | 0 | 0 | 201,753 | 201,753 |
| 1992 | 1993 | 0 | 0 | 594,000 | 594,000 |
| 1993 | 1994 | 4,817,929 | 0 | 550,700 | 5,368,629 |
| 1994 | 1995 | 2,334,264 | 0 | 289,500 | 2,623,764 |
| 1995 | 1996 | 2,691,311 | 6,138 | 572,350 | 3,269,799 |
| 1996 | 1997 | 3,038,171 | 0 | 96,500 | 3,134,671 |
| 1997 | 1998 | 0 | 0 | 437,950 | 437,950 |
| 1998 | 1999 | 0 | 0 | 0 | 0 |
| 1999 | 2000 | 0 | 0 | 0 | 0 |
| 2000 | 2001 | 2,743,374 | 0 | 0 | 2,743,374 |
| 2001 | 2002 | 0 | 0 | 0 | 0 |
| 2002 | 2003 | 0 | 0 | 49,500 | 49,500 |

aNumber of stocked fry updated from ADF&G Mark, Tag, and Age Laboratory website 20 June 2012.

bNumber of unmarked fry from incubation boxes from Eggers et al. 2010 (Table 1; FMS 10-05).

## Stock-Recruitment Analysis

Hierarchical stock-recruitment model were fit to the Chilkoot Lake stock-recruit data for the 1976 to 2010 brood years. The stock-recruit models are Ricker type (Ricker 1975) and hierarchal terms included escapement density and a first order autoregressive term. Three models were constructed: (1) linear, no density dependence due to escapement; (2) straight Ricker, density dependence due to escapement; and (3) autoregressive Ricker with the density dependence due to escapement and autoregressive terms included. The significance of the relative fit of the alternative models was evaluated using the likelihood ratio test (Hilborn and Mangel 1997).

Model 1, Linear;

 (4)

Model 2, Straight Ricker;

 (5)

Model 3, Autoregressive Ricker. The autoregressive Ricker model is the result of a first order (mean = 0, parameter autoregressive process where observations are linearly related to the prior year observation (c.f. Noakes et al 1987).

 (6)

## State Space Model

Chilkoot Lake sockeye salmon spawner-recruit data were analyzed using a Bayesian age-structured state-space model to assess the uncertainty introduced into the estimate of spawning size that produces maximum sustained yield (MSY) due to the following factors.

1. Late installation of the weir (between 18 June and 13 July) in years 1982, 1983, 1985, 1987, 1988, 1999, and 2001–2007, and early removal of the weir (between 28 September and 14 October) in years 1977, 1980, 1982, 1984, 1994, 1995, 2001, and 2003–2006. In years 1982, 2001, and 2003–2006 the weir was installed late *and* removed early. Therefore, the spawning escapement counts had to be expanded in years 1977, 1980, 1982, 1983, 1984, 1985, 1987, 1988, 1994, 1995, 1999, and 2001–2007.
2. Late installation of the DIDSON (27 June in 2008, 26 June in 2015, 24 June 2016) in years 2008, 2015, and 2016, and early removal (between 30 September and 7 October) in years 2012, 2013, and 2015. Therefore, the spawning escapement counts had to be expanded in years 2008, 2012, 2013, 2015, and 2016.
3. Multiple, overlapping methods of escapement enumeration; Weir counts (1976–2007) are reliable, but provide minimum estimates of escapement due to flow reversals, turbid water, and increased lowering of the boat gate allowing undetected fish passage. Therefore, weir counts are an index of escapement. Compared to the weir, the DIDSON (2008–2016) has the potential to provide highly accurate counts of fish; however, the DIDSON count is also considered a minimum estimate due to undetected fish passage at night, misidentification of species, and miscounts of fish. Mark–recapture estimates (1994–2016) may be greatly inflated, but may provide an index of escapement. Since the mark–recapture estimates are highly inflated and the weir counts are highly underestimated, the DIDSON escapement counts were treated as the ‘true’ counts and not as an index of escapement.
4. The weir was not operated from 1996 to 1998. Therefore, proportions of ages 2–8 were only available for the commercial harvest and weighted annual proportions by age (escapement and harvest combined) were considered unknown.

### Process Model

A hierarchical set of two stock-recruitment models were fit to the Chilkoot Lake stock-recruit data for calendar years 1976–2016. The stock-recruitment models were Ricker-type and hierarchical terms included density-dependent fry plants and a first order autoregressive (AR(1)) term. Returns *R* of Chilkoot lake sockeye salmon were modeled as a function of spawning escapement *S* in year *y* using a linearized Ricker (1954) spawner-recruit function with an AR lognormal process error with a lag of 1 year (Noakes et al. 1987),

. (3)

Returns for Model 2 were modeled using a linearized AR(1) Ricker (eq. 3), but an additional density-dependent fry term, *γ*, was included,

. (4)

Fry plants were from brood year *y* in year *y*+1 (Fy+1). In eq. 3 and 4, **is the productivity parameter, ** is the inverse capacity parameter, **is the AR lag-1 coefficient, and {} are the model residuals,

 (5)

In eq. 3, {} are independently and normally distributed process errors with standard deviation  Six initial returns (1970–1975) were modeled as draws from a common log normal distribution with parameters ln(*R*0) and  These returns were not linked to the escapement data in the spawner recruit relationship.

Age-at-maturity proportions  from year *y* and returning at ages 4–6 (ages 2–4 were combined and ages 6–8 were combined) were drawn from a common Dirichlet distribution that was implemented by generating independent random variables from the gamma distribution  and dividing each by their sum (Evans et al. 1993),

. (6)

Proportions of recruits at age, , (Gelman et al. 2004) were calculated as

, (7)

and implemented as a series of nested beta distributions, reflecting age-at-maturity central tendencies. The sum of the Dirichlet parameters, , is the inverse dispersion (*D*) of the Dirichlet distribution. A low value of *D* is reflective of a large amount of variability of age-at-maturity proportions *p* among brood years, whereas a high value of *D* indicates more consistency in *p* over time.

The abundance *N* of Chilkoot Lake sockeye salmon of age-*a* returning to spawn in calendar year *y* (*y* = 1976–2016) is the product of the age proportion scalar *p* and the total return (recruitment) *R* from year *y*-*a*,

. (8)

Total run abundance during calendar year *y* is the sum of abundance-at-age across ages,

. (9)

The spawning escapement count each calendar year, *Sy*, is the difference between total run abundance and the total District 15 Chilkoot Lake commercial harvest, *Hbelowy*,

 (10)

Annual commercial harvest was modeled as the product of the total run and annual harvest rate,

, (11)

drawn from a beta distribution (Appendix A.1).

### Observation Model

Observed data (Appendix A.2) included spawning escapement counts (DIDSON), indices of escapement (weir, mark–recapture), annual commercial harvest, and age compositions. For this analysis, we assume no unreported harvest of Chilkoot Lake sockeye salmon.

Estimated escapement counts from the DIDSON were

, (12)

where the  were normal (0, ) and

 (13)

Estimated annual commercial harvest was

, (14)

where the {} were normal (0, ) and the variances followed Equation 13. Two indices of escapement were available. Each comprised an independent measure of relative escapement,

, (15)

where subscript *i* indicates 1 of the 2 indices of escapement (weir or mark–recapture), *qi* is a factor of proportionality relating true escapement to index *Ii* ,and the are independently and normally distributed process errors with variance . Parameters *qi* and  were estimated from the data.

Harvest coefficients of variation were unavailable. Therefore, they were uniformly set to an arbitrarily high value of 0.20 so as not to overstate confidence in the harvest estimates. For the years when no temporal expansion of DIDSON or weir counts was necessary, the CV of the spawning escapement was set to an arbitrarily small value of 0.05. Fleischman et al. (2013) found that results from a similar analysis were not sensitive to arbitrary choices of weir count CVs. For years when DIDSON or weir escapement counts were expanded for either late installation *or* early removal, the CVs were set at 0.10 (Bednarski et al. 2017). For years when DIDSON or weir escapement counts were expanded for both late installation *and* early removal, the CVs were set at 0.20. The CVs for mark–recapture escapement counts were estimated as the standard error of the drainagewide point estimate divided by the drainagewide point estimate. These are standard output from the software program Stratified Population Analysis System (SPAS) that was used to analyze the mark–recapture data (Arnason et al. 1996). The CVs for 1994–1997 were unavailable. Therefore, they were uniformly set at 0.10.

For both annual commercial harvest and escapement samples separately, proportions of age 2–8 fish by return year were first converted to numbers by age based on the annual escapement and harvest numbers. Then, the numbers by age for annual escapement and annual harvest were combined for each age group (ages 2–8). Next, these combined numbers by age were converted to annual proportions by age, *q*(ob)*y,a*. This method basically weights the proportions by the escapement and harvest numbers (i.e., if harvest was higher, the proportions by age in the harvest received more weight). Since effective sample size could not be accurately calculated for escapement or harvest due to unknown variances, and key model results from state-space analyses of Pacific salmon are typically not sensitive to the choice of *nEy* (Fleischman and McKinley 2013), an arbitrarily small annual effective sample size of *nEy* =100 was used. After combining proportions of ages two through four and also combining ages six through eight, the weighted annual proportions by age were multiplied by 100,

 where  across all ages for each year, (16)

to calculate the age counts, *xy,a*. The age counts were assumed to have a multinomial distribution with order parameter *nEy*and proportion parameters,

, (17)

where  across all ages for each calendar year. The weir was not operated from 1996–1998. Therefore, proportions of ages 2–8 were only available for the commercial harvest and weighted annual proportions by age for the combined escapement and harvest data were considered unknown.

## Model Fitting

Model fitting involves finding the values of population parameters that can plausibly result in the observed data. Using the package RJAGS (Plummer 2016) within R (R Core Team 2016), Markov Chain Monte Carlo (MCMC) methods were employed to provide a more realistic assessment of uncertainty than is possible with traditional stock-recruit methods.

The significance of the fit of the alternative models (Model 1: l linearized AR(1) Ricker; Model 2: linearized AR(1) Ricker with density-dependent fry plants) were evaluated using the deviance information criterion (DIC; Spiegelhalter et al. 2002), a Bayesian version or generalization of the Akaike information criterion (AIC; Akaike 1973) and related to the Bayesian (or Schwarz) information criterion (BIC; Schwarz 1978). Similar to both the AIC and BIC, it trades off a measure of model adequacy against a measure of complexity. Models receiving a DIC within 1-2 of the ‘best’ model deserve consideration, while models within 3–7 of the ‘best’ model have considerably less support (Spiegelhalter et al. 2002).

### Prior Distributions

For all unknowns in the model, Bayesian analysis requires that prior probabilities be specified. Most prior distributions in this model were uninformative with a few exceptions (Table 4). Normal priors with mean 0, extremely large variances, and constrained to be positive were used for *β* (Millar 2002). Log transformed initial recruitments *R*1970*-R*1976 (those with no linked spawner abundance) were modeled as drawn from a common normal distribution with mean ln(*R*0) and variance . A flat prior on the standard deviation of log initial brood year returns, , caused computational disruptions during MCMC sampling so it was changed to a slightly informative inverse gamma prior. Fleischman et al. (2013) found that an informative prior on  may have a large effect on the posterior of  and the initial values of *Ry*, but negligible effects on key model quantities.

Table 4.–Prior distributions for model parameters. Where “Uniform” is in quotes, a normal distribution with mean 0 and large variance was used in the actual RJAGS code to prevent computational disruptions during MCMC sampling.

|  |  |  |
| --- | --- | --- |
| **Parameter** | **RJAGS coding** | **Prior** |
| ln(**) | lnalpha | ln(**) ~ “Uniform” (0,3) |
| ** | beta | **~ “Uniform” (0,∞) |
| *R* | sigma.R | ~gamma(0.001,0.001) |
| ** | phi | **~ “Uniform” (-0.98,0.98); |
| ** | log.resid.0 |  |
| *D* | D |  |
| ln(*R*0) | mean.log.R0 | ln(*R*0) ~ “Uniform” (∞,∞) |
| *R0* | sigma.R0 | ~gamma(0.001,0.001) |
| *γ* | gam | *γ*~ “Uniform” (-0.98,0.98) |
| ln(*R*1):ln(*R*7) | log.R[1:7] | ~Normal(ln(*R*0), ) |

### Sampling from the Posterior Distribution

MCMC methods were used to generate the joint posterior probabilities of the unknown quantities using the package RJAGS (Plummer 2016) with R. Three Markov chains were initiated. After a 10,000 sample burn-in period was discarded, 3,000 samples (1,000,000 iterations, thinned by 1000; 1000 samples per chain) MCMC updates were retained for analysis to estimate posterior medians, standard deviations, and percentiles. The diagnostic tools of the package RJAGS (Plummer 2016) such as time series and density plots, the Gelman Rubin convergence diagnostics (Brooks and Gelman 1998), autocorrelation plots, and Monte Carlo standard errors (e.g. MC error should be less than 5% of the sample standard deviation) (Toft et al. 2007) were used to assess mixing and convergence. No major problems were encountered. Interval estimates were constructed from the percentiles of the posterior distribution.

### Reference Points, Optimal Yield Profiles, Overfishing Profiles, Optimal Recruitment Profiles, and Sustained Yield

Reference points were calculated for each individual MCMC sample. Spawning abundance at maximum sustained yield (MSY), *S*MSY, was approximated by (Hilborn 1985),

 (18a)

and approximated based on Peterman et al. (2000),

 (18b)

where to correct for the difference between the median and the mean of a lognormal error distribution from an AR(1) process (Parken et al. 2006). Sustained yield at a specified level of *S* was obtained by subtracting spawning escapement from recruitment,

 (19)

Spawning escapement at peak return, *S*MSR, was calculated as 1/ and equilibrium spawning abundance (recruitment exactly replaces spawners) as,

. (20)

Harvest rate leading to MSY, *U*MSY, was approximated by (Hilborn 1985),

, (21a)

and approximated based on Peterman et al. (2000),

 (21b)

Optimal yield probabilities are the probabilities that a given level of spawning escapement (*S*) will produce average yields exceeding X% of MSY. These probabilities are created by calculating expected sustained yield, *YS*, at incremental levels of *S* (0 to 300,000 by 2,500) for each MCMC sample using equation 19, and then comparing *YS* with X% (80%, 90%) of the value of MSY for the sample. The proportion (*P*OY) of samples that fit the criteria: *YS* > X% of MSY is an estimate of the desired probability. Optimal yield profiles are plots of *P*OY versus *S* (Fleischman et al. 2013).

Overfishing probability profiles show the probability of overfishing the stock such that sustained yield is reduced to less than a fraction (80%, 90%) of MSY. To produce the overfishing probability profiles, expected sustained yield (eq. 19) at multiple incremental levels of *S* (0 to 300,000 by 2,500) are calculated for each MCMC sample. Then, the number of MCMC samples for which *YS*is less than X% of MSY and *S* is less than *S*MSY is tabulated. Overfishing probability profiles are then a plot of the fraction of samples in which this condition occurred versus *S* (Bernard and Jones III 2010).

Optimal recruitment profiles are the probabilities that a given spawning escapement (*S*) will produce average recruitments (*R*) exceeding X% (80%, 90%) of maximum sustained recruitment (MSR). These probabilities are created by calculating *R* from

 (23)

at incremental levels of *S* (0 to 300,000 by 2,500) for each MCMC sample, then comparing *R* with X% of the value of MSR for that sample. The proportion *PR* of samples in which *R* exceeded X% of MSR is an estimate of the desired probability. Optimal recruitment profiles are then a plot of *PR* versus *S* (Fleischman et al. 2013).

Expected sustained yield is the number of fish in the expected recruitment over and above that needed to replace the spawners (Fleischman et al. 2011).

# Results

Appendix A2 summarizes empirical (data-based) estimates of harvest, age composition, escapement, and escapement indices for Chilkoot Lake sockeye salmon.

### Model Comparison

A hierarchical set of two stock-recruitment models were fit to the Chilkoot Lake stock-recruit data for brood years 1976–2016. The stock-recruitment models were Ricker-type and hierarchical terms included a first order autoregressive term. The effect of the fry plant term (*γ* = -5.81E-08; Table 5) was to correct the increased production due to the fry plants and to provide an unbiased estimate of the wild stock MSY escapement goal. The best model based on the fit criteria (i.e., minimum DIC) was the autoregressive Ricker with fry plants, but the difference between the best fit model fit and the autoregressive Ricker model fit was negligible (∆0.31) and the 95% credibility intervals on the density-dependent fry term, *γ*, included zero (-2.74E-07, 1.69E-07) and thus the term was non-significant. Therefore, the autoregressive Ricker model was considered the most meaningful biological model.

Table 5.–Results of model fit to the escapement-recruit data for brood years 1976 to 2010. Estimated parameters, reference points, and measures of fit (DIC). The lower and upper bound on the *S*MSY and *S*MSY(Peterman) reference points define the 2.5th and 97.5th percentiles which represent the 95% credibility intervals for the parameters.

|  |  |
| --- | --- |
| Parameters | Autoregressive Ricker |
| *α* | 2.89 |
| *β* | 5.11E-06 |
| ** | 0.49 |
| *S*MSY | 101,425 |
| Lower | 65,484 |
| Upper | 308,889 |
| *S*MSY(Peterman) | 101,626 |
| Lower | 65,520 |
| Upper | 311,043 |
| Harvest Rate | 0.73 |
| Harvest Rate Peterman | 0.73 |
| DIC | 3090.20 |

### Abundance, Time-Varying Productivity, Harvest Rates, and Age-at-Maturity

Reconstructed total run abundance (*N*) for the autoregressive Ricker model had CVs from 5% to 14% (Figure 3; Table 6). The years with higher uncertainty corresponded to years with missing

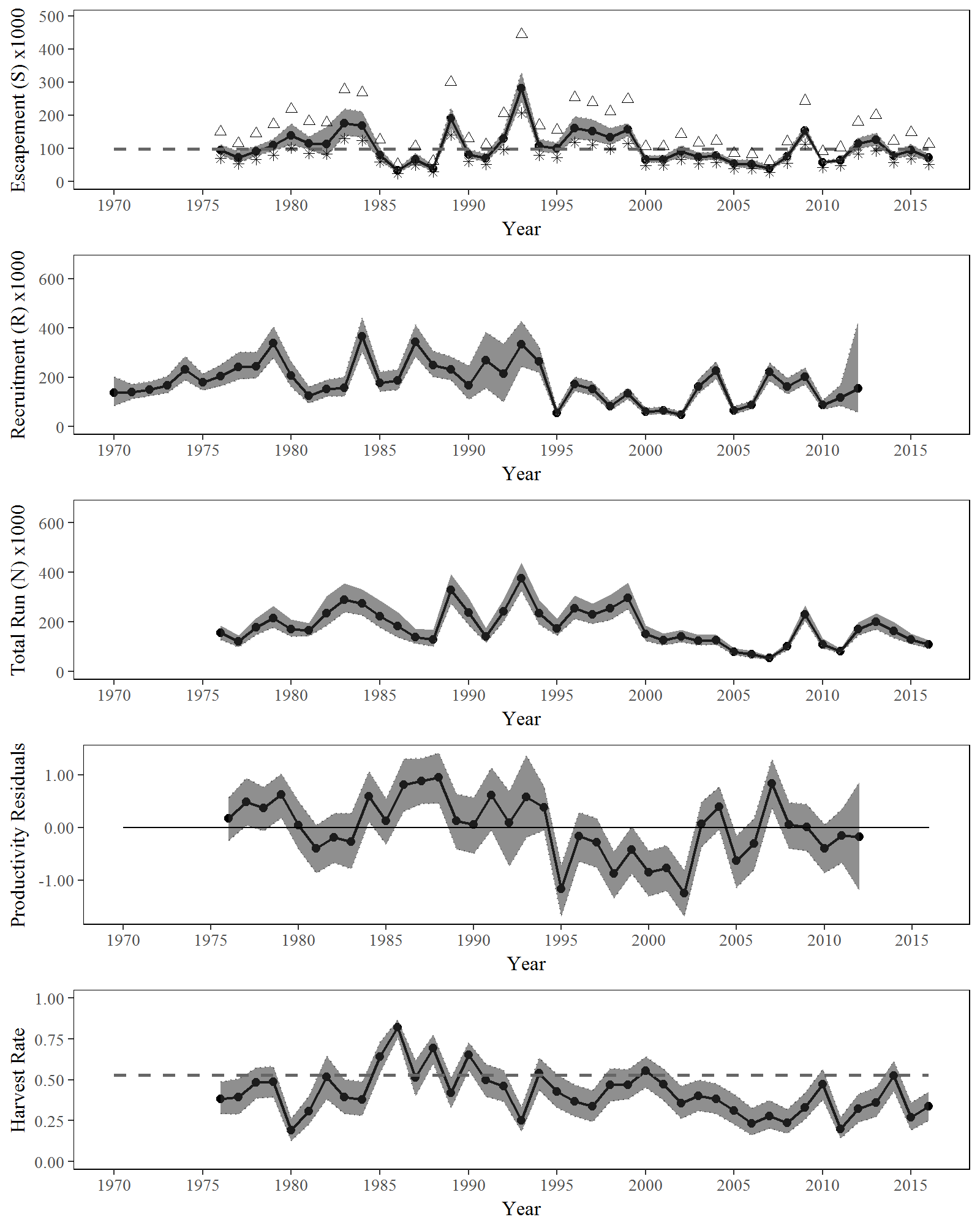


Figure 3.–Point estimates (posterior medians; solid lines) and 95% credibility intervals (shaded areas) of spawning escapement with the two indices of escapement (weir (stars); mark–recapture (open triangles)), recruitment by brood year, total run abundance, Ricker productivity residuals, and harvest rate from a state-space model of Chilkoot Lake sockeye salmon, 1976–2016. Posterior medians of *S*MSY and *U*MSY are plotted as dashed horizontal reference lines.

Table 6.–Annual abundance estimates for Chilkoot Lake sockeye salmon obtained by fitting a state-space model to data for calendar years 1976–2016. Point estimates are posterior medians and CVs are posterior standard deviations divided by posterior means. Recruitment values are listed by brood year.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
|  |  |  |  | Coefficients of Variations | |  |
| Year | Total Run | Escapement | Return | Total Run | Escapement | Return |
| 1970 |  |  | 137,316 |  |  | 0.22 |
| 1971 |  |  | 138,731 |  |  | 0.11 |
| 1972 |  |  | 149,647 |  |  | 0.10 |
| 1973 |  |  | 166,317 |  |  | 0.10 |
| 1974 |  |  | 230,729 |  |  | 0.10 |
| 1975 |  |  | 178,517 |  |  | 0.10 |
| 1976 | 154,759 | 94,767 | 203,002 | 0.09 | 0.08 | 0.11 |
| 1977 | 120,150 | 72,454 | 240,751 | 0.10 | 0.12 | 0.12 |
| 1978 | 177,258 | 91,665 | 243,163 | 0.10 | 0.08 | 0.11 |
| 1979 | 213,916 | 109,033 | 337,173 | 0.10 | 0.08 | 0.10 |
| 1980 | 171,241 | 138,874 | 205,394 | 0.10 | 0.11 | 0.12 |
| 1981 | 166,555 | 114,918 | 125,111 | 0.08 | 0.08 | 0.13 |
| 1982 | 235,682 | 112,566 | 152,666 | 0.13 | 0.19 | 0.11 |
| 1983 | 289,335 | 175,171 | 155,779 | 0.10 | 0.11 | 0.13 |
| 1984 | 273,751 | 169,155 | 365,918 | 0.10 | 0.11 | 0.09 |
| 1985 | 223,643 | 80,025 | 177,253 | 0.12 | 0.12 | 0.12 |
| 1986 | 182,155 | 32,763 | 187,376 | 0.14 | 0.08 | 0.11 |
| 1987 | 138,713 | 67,242 | 343,729 | 0.11 | 0.12 | 0.09 |
| 1988 | 129,464 | 39,289 | 248,168 | 0.13 | 0.12 | 0.11 |
| 1989 | 328,466 | 190,175 | 231,251 | 0.09 | 0.08 | 0.10 |
| 1990 | 236,104 | 82,037 | 166,363 | 0.12 | 0.08 | 0.21 |
| 1991 | 141,265 | 70,541 | 268,761 | 0.11 | 0.08 | 0.22 |
| 1992 | 241,875 | 129,999 | 213,594 | 0.10 | 0.08 | 0.28 |
| 1993 | 376,584 | 281,305 | 333,692 | 0.08 | 0.08 | 0.14 |
| 1994 | 234,189 | 107,365 | 263,850 | 0.11 | 0.09 | 0.10 |
| 1995 | 174,126 | 99,241 | 54,392 | 0.10 | 0.08 | 0.16 |
| 1996 | 255,615 | 161,020 | 170,871 | 0.10 | 0.11 | 0.09 |
| 1997 | 228,785 | 151,191 | 151,164 | 0.09 | 0.11 | 0.09 |
| 1998 | 253,565 | 133,811 | 81,823 | 0.10 | 0.10 | 0.12 |
| 1999 | 296,766 | 156,537 | 133,277 | 0.09 | 0.06 | 0.08 |
| 2000 | 149,615 | 66,770 | 59,773 | 0.11 | 0.07 | 0.11 |
| 2001 | 126,445 | 66,897 | 65,214 | 0.09 | 0.08 | 0.10 |
| 2002 | 140,668 | 90,054 | 48,135 | 0.09 | 0.10 | 0.11 |
| 2003 | 124,193 | 74,104 | 160,581 | 0.09 | 0.08 | 0.09 |
| 2004 | 125,953 | 77,162 | 225,809 | 0.08 | 0.07 | 0.08 |
| 2005 | 78,435 | 53,820 | 64,638 | 0.09 | 0.10 | 0.12 |
| 2006 | 68,537 | 52,341 | 87,719 | 0.10 | 0.12 | 0.09 |
| 2007 | 54,114 | 38,855 | 220,408 | 0.08 | 0.08 | 0.08 |
| 2008 | 100,740 | 76,643 | 161,408 | 0.07 | 0.07 | 0.10 |
| 2009 | 230,761 | 154,202 | 201,367 | 0.07 | 0.05 | 0.09 |
| 2010 | 109,848 | 57,719 | 86,419 | 0.09 | 0.05 | 0.11 |
| 2011 | 81,821 | 65,333 | 117,568 | 0.05 | 0.04 | 0.18 |
| 2012 | 169,313 | 114,024 | 154,752 | 0.08 | 0.07 | 0.56 |
| 2013 | 200,024 | 126,732 |  | 0.08 | 0.07 |  |
| 2014 | 164,112 | 77,737 |  | 0.10 | 0.05 |  |
| 2015 | 129,403 | 93,986 |  | 0.08 | 0.09 |  |
| 2016 | 108,283 | 71,703 |  | 0.08 | 0.07 |  |

escapement data (DIDSON; 1976–2007), missing escapement indices (weir or mark–recapture), and/or missing age composition data (1996–1998; Table 6; Appendix A2). Excluding the first initial returns, reconstructed brood year recruitment had CVs from 8% to 56%. The Ricker recruitment residuals (productivity residuals) in Figure 3 are deviations in recruitment from that predicted by the Ricker S–R relationship, reflecting time-varying changes in productivity after controlling for density-dependent effects. Productivity residuals were spread around 0 across years, indicating a good model fit and *σ*R was 0.55 (95% CI: 0.43–0.74) (Table 7). Median harvest rates (*U*MSY) ranged from 0.37 to 0.71 (Figure 3) and ranged from 0.37 to 0.70 for *U*MSY(Peterman). Chilkoot Lake sockeye salmon matured at ages 2– 4 (mean **4; 1%–8%), age-5 (mean **5; 36%–72%), and ages 6–8 (mean **6; 23% to 63%). These proportions have fluctuated moderately from brood year to brood year (Figure 4; top panel). Chilkoot Lake sockeye salmon are mainly composed of age–5 (1.3, 2.2) and age–6 (2.3) fish (Table 8). Age compositions have also fluctuated from year to year (Figure 4; middle and bottom panel).

Table 7.–State-space model parameter estimates for Chilkoot Lake for calendar years 1976–2016. Posterior medians are point estimates; 2.5th and 97.5th percentiles define 95% credibility intervals for the parameters. Parameter definitions are in the *Methods* section. Point estimates are posterior medians and CVs are posterior standard deviations divided by posterior means.

|  |  |  |  |  |
| --- | --- | --- | --- | --- |
| Parameter | 2.5th percentile | Median | 97.5th percentile | CV |
| ** | 1.81 | 3.02 | 5.21 | 0.28 |
| ln(**) | 0.59 | 1.11 | 1.65 | 0.24 |
| ** | 1.90E-06 | 5.45E-06 | 9.24E-06 | 0.35 |
| ** | 0.07 | 0.42 | 0.76 | 0.42 |
| **R | 0.43 | 0.55 | 0.74 | 0.14 |
| *S*EQ | 165,943 | 239,069 | 555,984 | 0.42 |
| *S*MSR | 108,227 | 183,490 | 527,649 | 0.58 |
| *S*MSY | 66,413 | 97,275 | 231,148 | 0.43 |
| *U*MSY | 0.37 | 0.53 | 0.71 | 0.16 |
| *S*MSY(Peterman) | 66,447 | 97,497 | 232,408 | 0.43 |
| *U*MSY(Peterman) | 0.37 | 0.54 | 0.70 | 0.16 |
| D | 20.50 | 32.45 | 51.78 | 0.24 |
| ** | 0.03 | 0.04 | 0.06 | 0.15 |
| ** | 0.51 | 0.54 | 0.57 | 0.03 |
| ** | 0.38 | 0.41 | 0.45 | 0.04 |
| *q*m-r | 1.46 | 1.58 | 1.72 | 0.04 |
| *q*weir | 0.65 | 0.74 | 0.83 | 0.06 |

Note: The CVs for the reference points *S*EQ, *S*MSR, *S*MSY, and *S*MSY(Peterman) were calculated as (97.5th percentile-2.5th percentile)/3.92 / posterior median point estimate. If the posterior median is approximately normal, then the lower and upper bound of the 95% credibility are both ~1.96 x standard errors from the median point estimate.

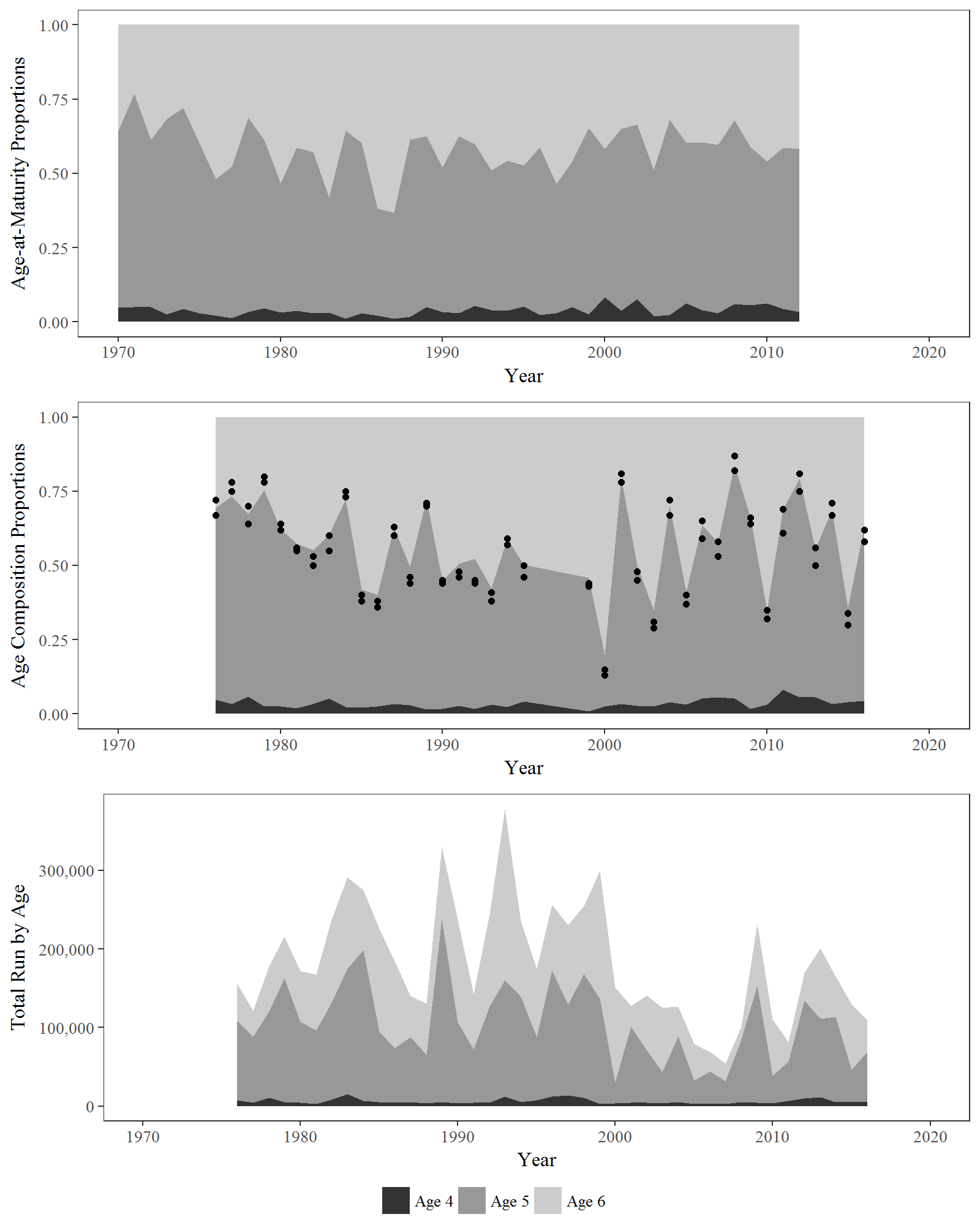


Figure 4.–Estimated age-at-maturity proportions by brood year (1970–2012; top), age composition proportions by calendar year (1976–2016; middle), and total run by age (bottom), from a state-space model fitted to data from Chilkoot Lake sockeye salmon. Top and middle figures are area graphs in which distance between lines represent age proportions. Dots in the middle plot are data-based estimates of age composition from Appendix A2.

Table 8.–Total run abundance by age-class obtained by fitting a state-space model to data from Chilkoot Lake sockeye salmon for calendar years 1976–2016. Point estimates are posterior medians and CVs are posterior standard deviations divided by posterior means.

|  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- |
|  |  |  |  | Coefficients of Variations | | |
| Year | Age-4 | Age-5 | Age-6 | Age-4 | Age-5 | Age-6 |
| 1976 | 7,220 | 99,708 | 47,244 | 0.40 | 0.11 | 0.17 |
| 1977 | 3,760 | 84,142 | 31,817 | 0.48 | 0.12 | 0.18 |
| 1978 | 9,859 | 109,095 | 57,578 | 0.38 | 0.12 | 0.16 |
| 1979 | 4,889 | 155,475 | 52,417 | 0.54 | 0.12 | 0.18 |
| 1980 | 4,068 | 101,874 | 64,302 | 0.55 | 0.12 | 0.15 |
| 1981 | 2,757 | 92,507 | 70,848 | 0.64 | 0.11 | 0.13 |
| 1982 | 7,500 | 121,748 | 104,937 | 0.50 | 0.15 | 0.17 |
| 1983 | 14,506 | 158,817 | 115,008 | 0.40 | 0.13 | 0.15 |
| 1984 | 5,745 | 191,204 | 75,516 | 0.56 | 0.12 | 0.17 |
| 1985 | 4,313 | 88,984 | 129,554 | 0.56 | 0.17 | 0.15 |
| 1986 | 4,136 | 68,373 | 109,233 | 0.56 | 0.18 | 0.16 |
| 1987 | 4,431 | 82,265 | 51,177 | 0.48 | 0.13 | 0.16 |
| 1988 | 3,377 | 60,297 | 65,140 | 0.55 | 0.16 | 0.16 |
| 1989 | 4,546 | 231,198 | 89,939 | 0.67 | 0.11 | 0.17 |
| 1990 | 3,420 | 101,139 | 129,709 | 0.66 | 0.16 | 0.14 |
| 1991 | 3,626 | 67,296 | 69,874 | 0.55 | 0.14 | 0.14 |
| 1992 | 3,800 | 121,830 | 115,242 | 0.66 | 0.14 | 0.14 |
| 1993 | 10,757 | 147,488 | 216,630 | 0.51 | 0.14 | 0.12 |
| 1994 | 4,933 | 132,262 | 95,515 | 0.55 | 0.14 | 0.15 |
| 1995 | 6,943 | 80,007 | 86,326 | 0.44 | 0.14 | 0.13 |
| 1996 | 8,689 | 161,579 | 79,936 | 0.99 | 0.24 | 0.36 |
| 1997 | 10,214 | 114,822 | 99,112 | 0.87 | 0.29 | 0.31 |
| 1998 | 7,859 | 158,142 | 83,642 | 0.89 | 0.22 | 0.35 |
| 1999 | 2,454 | 133,142 | 160,372 | 0.69 | 0.13 | 0.13 |
| 2000 | 3,516 | 25,527 | 119,870 | 0.55 | 0.21 | 0.12 |
| 2001 | 4,069 | 96,370 | 25,503 | 0.48 | 0.11 | 0.20 |
| 2002 | 3,731 | 65,867 | 70,101 | 0.48 | 0.14 | 0.13 |
| 2003 | 2,963 | 39,944 | 80,775 | 0.55 | 0.16 | 0.11 |
| 2004 | 4,714 | 83,341 | 37,350 | 0.40 | 0.11 | 0.16 |
| 2005 | 2,257 | 29,698 | 46,239 | 0.49 | 0.14 | 0.12 |
| 2006 | 3,496 | 39,898 | 24,864 | 0.37 | 0.13 | 0.16 |
| 2007 | 2,926 | 28,264 | 22,683 | 0.38 | 0.12 | 0.13 |
| 2008 | 5,128 | 79,296 | 16,056 | 0.39 | 0.09 | 0.20 |
| 2009 | 3,658 | 148,245 | 78,266 | 0.55 | 0.10 | 0.14 |
| 2010 | 3,171 | 34,826 | 71,468 | 0.48 | 0.16 | 0.12 |
| 2011 | 6,493 | 49,572 | 25,466 | 0.32 | 0.09 | 0.14 |
| 2012 | 9,301 | 124,695 | 34,538 | 0.36 | 0.10 | 0.19 |
| 2013 | 10,929 | 99,599 | 88,768 | 0.37 | 0.12 | 0.13 |
| 2014 | 5,145 | 106,674 | 51,404 | 0.44 | 0.12 | 0.17 |
| 2015 | 4,786 | 41,066 | 82,783 | 0.43 | 0.15 | 0.11 |
| 2016 | 4,546 | 63,641 | 39,569 | 0.44 | 0.11 | 0.14 |

### Stock Productivity, Capacity, and Yield

The Ricker stock recruit relationships derived from the age-structured state-space model fitted to escapement, harvest, and age composition data are variable. Results take into account measurement error in both *S* and *R* as depicted by the error bars in Figure 5, which weight the individual data pairs depending on how precisely they were estimated. Some of the plausible relationships (Figure 5; light grey lines) vary greatly from the posterior medians of ln(**) and **Figure 5; dark dashed line, but most are not substantially different from the median estimates*.* The median estimate of ln(*α*) was 1.11 (95% CI: 0.59–1.65), corresponding to *α* = 3.02 (95% CI: 1.81–5.21; Table 7) and the median estimate of the density dependent parameter *β* was 5.45 × 10-6 (95% CI: 1.90× 10-6–9.24 × 10-6). Uncertainty about *α* (CV = 0.28; Table 7) is evident in the extent to which the plausible S–R relationships differ with respect to their slope at *S* = 0 (Figure 5). Similarly, uncertainty about *β* is reflected in variability in the values of *S* leading to maximum recruitment *S*MSR= 1/*β,* and uncertainty about equilibrium abundance, *SEQ,* is reflected by variability in the values of *S* where the curves intersect the replacement line. The estimated AR(1) parameter **was 0.42 (95% CI: 0.07–0.76), suggesting serial correlation in residuals.

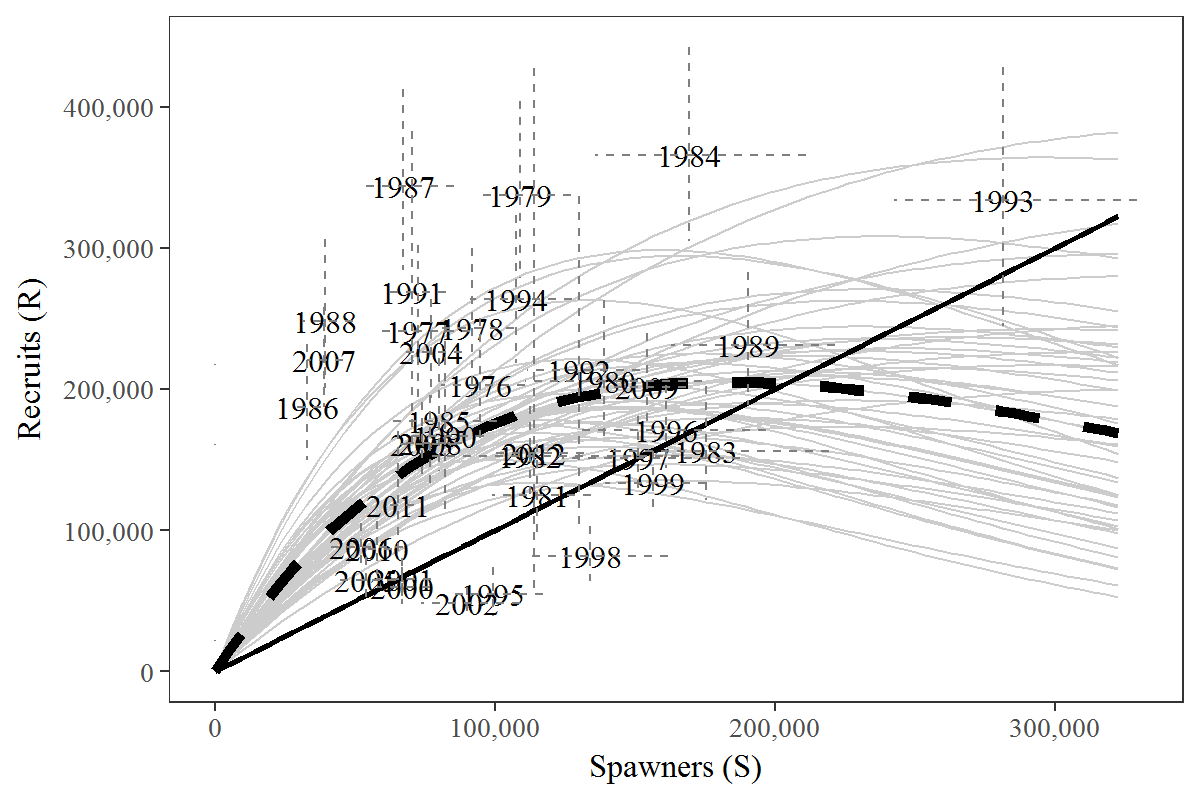


Figure 5.–Plausible spawner-recruit relationships for Chilkat Lake sockeye salmon as derived from an age-structured state-space model fitted to abundance, harvest, and age data for calendar years 1976–2016. Posterior medians of *R* and *S* are plotted as brood year labels with 95% credibility intervals plotted as dashed lines. The heavy dashed line is the Ricker relationship constructed from ln(*α*) and *β* posterior medians. Ricker relationships are also plotted (light grey lines) for 50 paired values of ln(*α*) and *β* sampled from the posterior probability distribution, representing plausible Ricker relationships that could have generated the observed data. Recruits replace spawners on the solid diagonal line.

Posterior medians of escapement leading to maximum sustained yield, *S*MSY and *S*MSY(Peterman),were 97,275 (95% CI: 66,413–231,148) and 97,497 (95% CI: 66,447–232,408), respectively (Figure 6). Given the diversity of plausible S–R relationships (Figure 5), it is important to choose

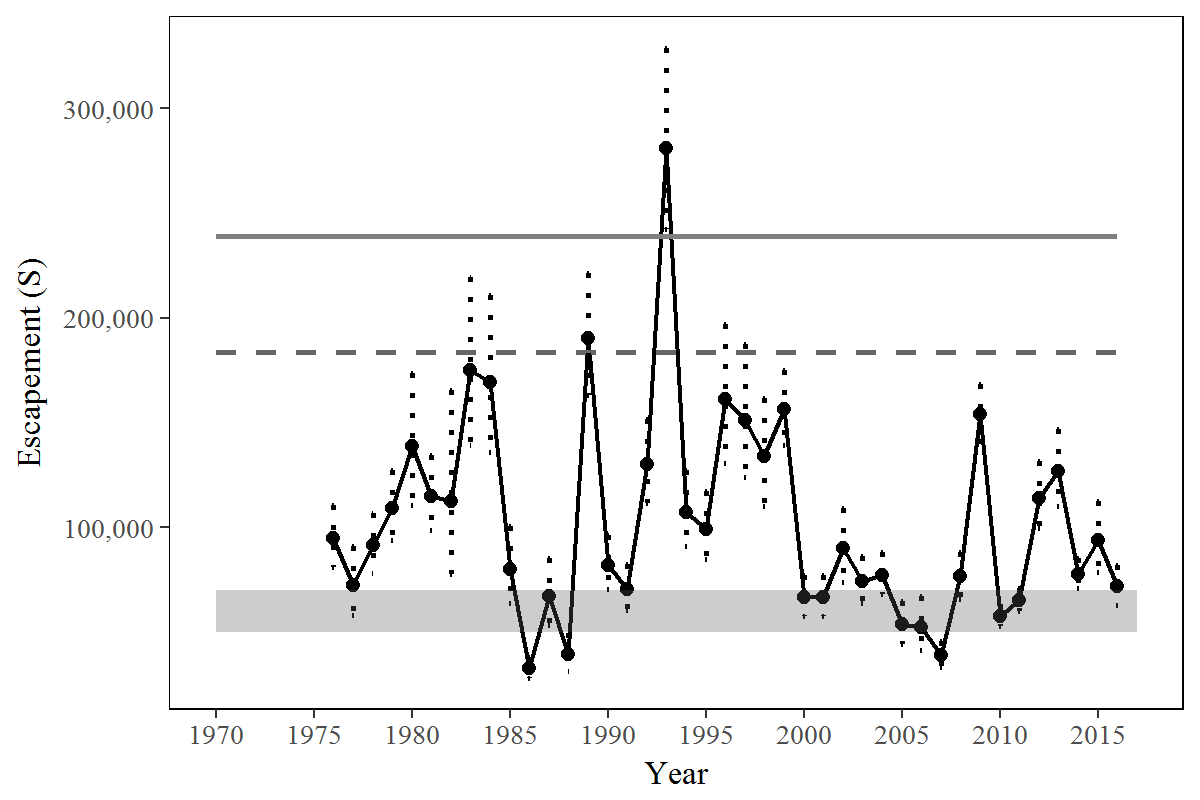


Figure 6.– Historical estimates of escapement and 95% credibility intervals (vertical dotted lines) for sockeye salmon obtained by fitting a state-space model to Chilkoot Lake sockeye salmon data, 1976–2016. Shaded areas bracket the recommended goal ranges. Posterior medians of *S*MSR (dotted line), and *S*EQ (solid line) are plotted as horizontal reference lines.

an escapement goal that is robust to this uncertainty rather than one tailored solely to the median S-R relationship. To address this uncertainty, the success or failure of a given number of spawners to achieve biological reference points across plausible S-R relationships are tallied to create optimal recruitment profiles (Figure 7; top panel), optimal yield profiles (Figure 7; middle panel), and overfishing profiles (Figure 7; bottom panel). Optimal recruitment profiles are the probabilities that a given spawning escapement will produce average recruitments exceeding 80% or 90% of maximum sustained recruitment. The profiles, which are highest near *S*MSR= 183,490 (Table 7), display the probability of achieving 80% and 90% of MSRfor specified levels of escapement. Optimal yield profiles show the probability of a given number of spawners achieving 80% and 90% of MSY. These probabilities, which are highest near *S*MSY, can be used to quantify the yield performance of prospective escapement goals (Figure 7; shaded areas), taking into consideration all of the uncertainty about the true abundance and productivity of the stock. Overfishing profiles show the probability that sustained yield would be reduced to less than 80% or 90% of MSYby fishing too hard and supplying too few spawners.

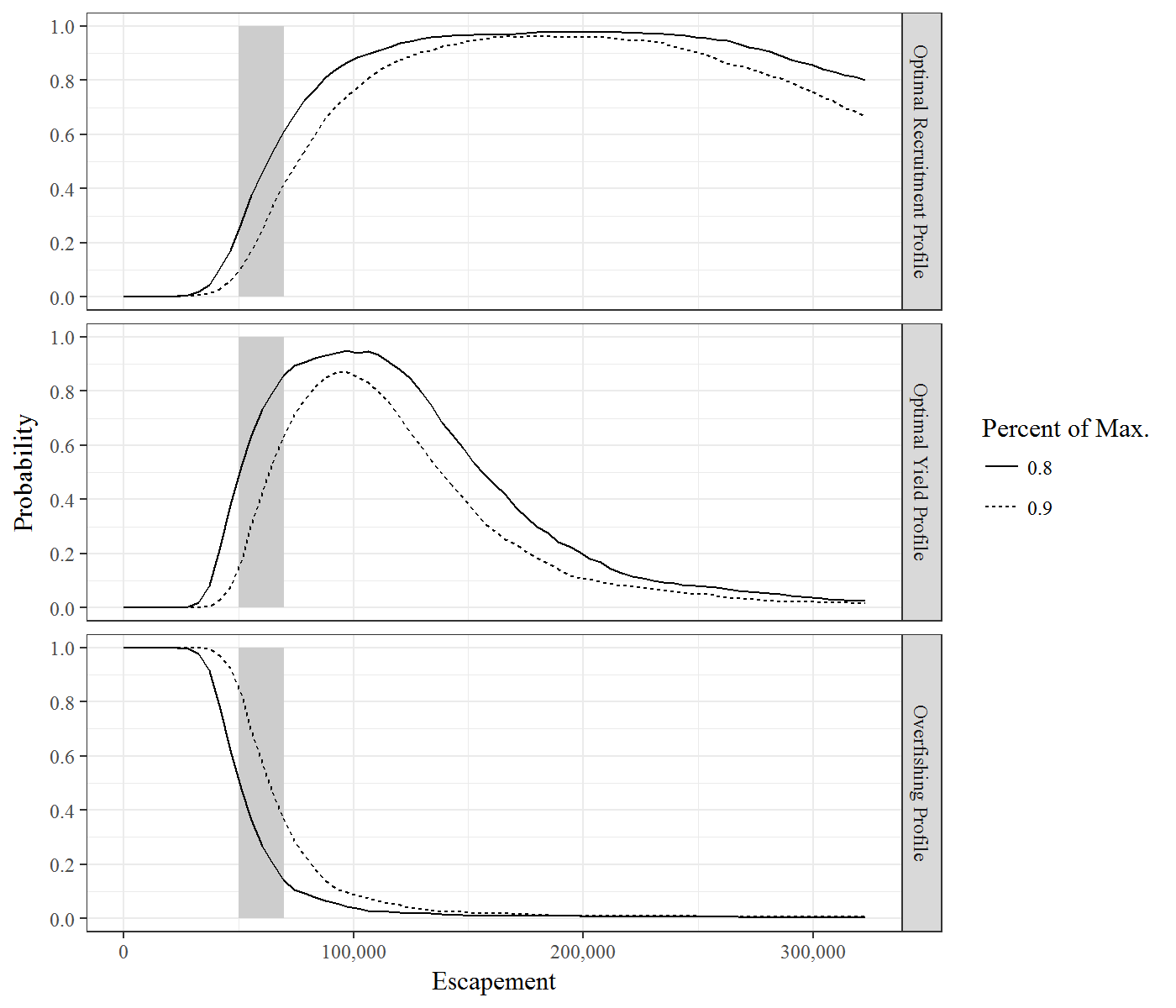


Figure .– Optimal recruitment profiles (ORPs), optimal yield profiles (OYPs), and overfishing profiles (OFPs) for Chilkoot Lake sockeye salmon. OYPs and ORPs show probability that a specified spawning abundance will result in specified fractions (80% and 90% line) of maximum sustained yield or maximum recruitment. OFPs show the probability that reducing escapement to a specified spawning abundance will result in less than specified fractions of maximum sustained yield. The shaded region shows the existing escapement goal range of 38,000-86,000.

Expected sustained yield or the numbers of fish over and above those necessary to replace spawners averaged over the brood years 1970–2012 is maximized near *S*MSY (Figure 8).

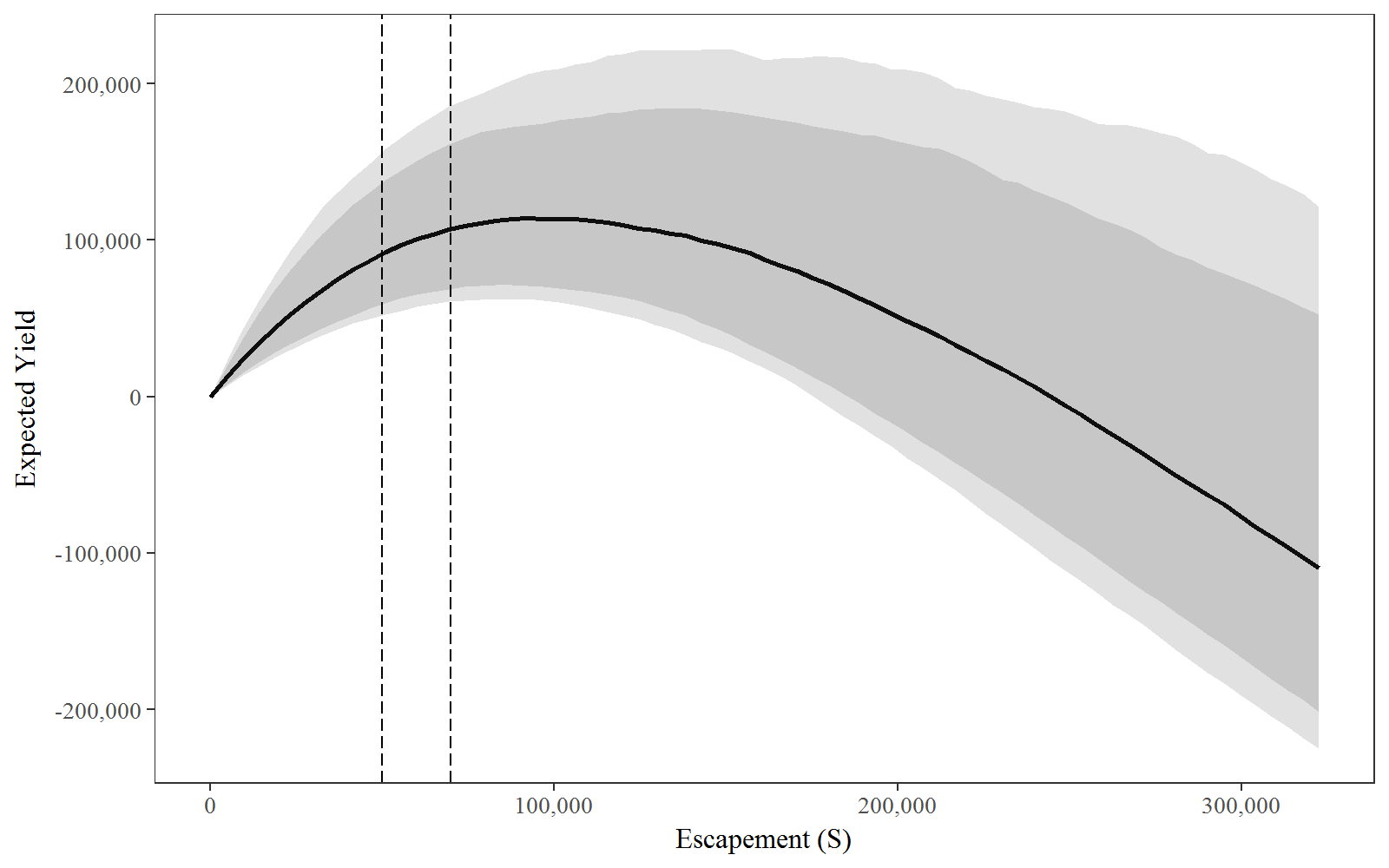


Figure 8.– Expected sustained yield (solid black line) and 90% and 95% credibility intervals (shaded areas) versus spawning escapement for Chilkoot Lake sockeye salmon. Dotted vertical lines bracket the recommended escapement goal range of ….

# Discussion

Based on the autoregressive Ricker model Eggers et al. (2009) recommended an SEG for Chilkoot Lake sockeye salmon of 38,000 to 86,000 spawners per year to be assessed with a weir at the Chilkoot River weir site. This goal range was the escapement range that produced 90% MSY as determined by the autoregressive Ricker model for the brood years 1976 to 2003 stock-recruit data This study was similar in methodology to Eggers et al. (2010) in that a set of hierarchical stock-recruitment models that incorporated a first order autoregressive term and fry plant term were constructed and model comparisons were accomplished through a fit criteria. While Eggers et al. (2010) utilized the traditional stock-recruit analysis, a Bayesian model approach was employed in this study. In traditional stock recruit analysis, independence of individual quantities of spawners (*S*) and recruits (*R*) is assumed, and missing data must be imputed before the model is run. One advantage of the Bayesian state-space model is that missing data are no longer an issue. By correctly specifying annual age-structure in the Bayesian state-space model, missing data, like parameters, can be represented as unknown quantities for which posterior samples are generated. Additional uncertainty then flows through to the remaining model parameters as appropriate. Along with overcoming the issue of missing data, another advantage of the Bayesian age-structured state space model over traditional stock recruit methods is on obtaining good quality estimates of spawning abundance at maximum sustained yield(*S*MSY) in regards to bias reduction and interval coverage (Su and Peterman 2012).

Within the traditional framework, Eggers et al. (2010) was forced to rely on one escapement enumeration method per year, although overlapping escapement enumeration methods were available. Eggers et al. (2010) regressed mark-recapture estimates against weir counts for years with paired estimates, so that weir counts could be expanded to total escapement during years when mark-recapture experiments were not conducted. Thus, an uninterrupted time series of escapement was created for analysis. Within the Bayesian analysis state-space framework, we were able to incorporate multiple, overlapping methods of escapement enumeration within the models with allowances for missing data. Weir and mark-recapture data were considered independent measures of relative escapement that likely under-estimated and over-estimated escapement, respectively, while the DIDSON counts were treated as ‘true’ counts.

In addition to the limitations of missing data and multiple sources of data, traditional stock-recruit analysis does not fully consider the complexity introduced by age structure. Sockeye salmon return at multiple ages to spawn and maturity schedules can differ between cohorts. Estimates of age composition of the total run, weighted by the relative abundance of each component (harvest, escapement), are incorporated into the state-space model as data input. The model then predicts maturity-at-age and age composition of the total run.

Bayesian age-structured state-space models are becoming increasing common for the analysis of escapement goal ranges for Pacific salmon in Alaska (Fleischman and Reimer 2017; Hamazaki et al. (2012); Fleischman and McKinley (2013)). Sockeye salmon escapement goals were assessed in a Bayesian framework for Buskin River, Kodiak Island (Schmidt and Evans 2010), Speel Lake (Heinl et al. 2014), and the transboundary Alsek River and one of its tributaries, the Klukshu River (Eggers and Bernard 2011). A sustainable escapement goal (SEG) range of 5,000-8,000 would ensure sustained yield is within 90% of *S*MSY with 90% probability for the Buskin River. Optimum yield profiles and overfishing profiles showed that a biological escapement goal (BEG) range of 24,000 to 33,500 for the Alsek River has a 90–96% chance of attaining optimum yield escapements and that a biological escapement goal range of 7,500 to 11,000 has a 79–90% chance of attaining optimum yield for the Klukshu River. An SEG of 4,000–9,000 fish for Speel Lake sockeye salmon is based on the range of escapements estimated to provide greater than 70–80% of *S*MSY.

In Alaska, most salmon BEGs are developed using Ricker spawner-recruit models (Ricker 1954), and by definition in the *Policy for the Management of Sustainable Salmon Fisheries* (5AAC 39.222), BEG ranges are estimates of the number of spawners that provide the greatest potential for maximum sustained yield (*S*MSY). Eggers (1993) suggests that an escapement goal range from 0.8 to 1.6 times the *S*MSY will enable managers more flexibility to protect weak stocks and maintain sustainable catch levels of dominant stocks to within 90% of *S*MSY. Based on Eggers (2009), the escapement goal range of Chilkoot Lake sockeye salmon to the nearest 100 fish would be would be an SEG of 38,000 to 78,000. Eggers et al. (2009) recommended an SEG instead of a BEG due to a relatively high level of uncertainty of the weir counts. Across the state, BEGs that were set based on spawner-recruit analyses had an average lower bound escapement goal range that was 0.68 times *S*MSY and an upper bound escapement foal range that was 1.40 times *S*MSY (Appendix C).

# Escapement Goal Recommendation

At this time we recommend that the existing escapement goal for Chilkoot Lake sockeye salmon remain unchanged. Results from our revised analysis that used updated data in a Bayesian framework suggest that the existing goal of 38,000–86,000 spawners are quite close to the range of spawners that are likely to result in a high probability (>70%) of achieving at least 90% of MSY.

Since 2012 there have been large escapements and the resulting recruits from these large brood year escapement will provide some good contrast……Thus, we suggest keeping the escapement goal unchanged until these recruits can be enumerated and included in a revised spawner-recruitment model.

Finally, as has been done for other stocks in recent years, we recommend using an age-structure modeling approach future analyses (Fleischman et al. 2013, Miller et al. in prep,…)…

# Acknowledgments

We would like to thank the many technicians and biologists who have helped to monitor and assess Chilkoot Lake sockeye salmon over the years, including those who have collected and read scales, operated weir and associated escapement projects.

SPA, genetics, weir, mark-recapture,

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# APPENDICES

Appendix .–RJAGS model code for the Bayesian MCMC analysis of the Chilkoot Lake sockeye salmon model, 1976–2016 can be found at the ADF&G GitHub site, located here: <https://github.com/commfish/AlaskaSalmon>. Please contact the authors of this report if you have problems opening this link or have questions or suggestion regarding the analysis.

Appendix B.–RJAGS data objects for the Bayesian MCMC statistical analysis of the Chilkoot Lake sockeye salmon data run reconstruction model, 1976–2016. The multinomial age counts (x) may not sum exactly to the effective sample size of 100 due to rounding. Y is the number of calendar years, A is the number of age classes, and C is the number of cohorts represented in the data (C=Y+A-1). In the table, ‘DS’ are the DIDSON escapement counts with associated coefficient of variations (DS.cv), ‘weir’ are the weir escapement counts with associated coefficient of variations (weir.cv), ‘mr’ are the mark–recapture escapement counts with the associated coefficient of variations (mr.cv), and ‘Hbelow’ is the total harvest with associated coefficient of variations (Hbelow.cv).

|  |  |  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Year | DS | DS.cv | mr | mr.cv | weir | weir.cv | Hbelow | Hbelow.cv | x4 | x5 | x6 |
| 1976 | NA | 0.90 | NA | 0.90 | 69,729 | 0.05 | 58,765 | 0.20 | 5 | 67 | 27 |
| 1977 | NA | 0.90 | NA | 0.90 | 50,363 | 0.10 | 41,477 | 0.20 | 3 | 75 | 22 |
| 1978 | NA | 0.90 | NA | 0.90 | 67,528 | 0.05 | 89,558 | 0.20 | 6 | 64 | 30 |
| 1979 | NA | 0.90 | NA | 0.90 | 80,588 | 0.05 | 115,995 | 0.20 | 2 | 78 | 20 |
| 1980 | NA | 0.90 | NA | 0.90 | 101,135 | 0.10 | 31,267 | 0.20 | 2 | 62 | 36 |
| 1981 | NA | 0.90 | NA | 0.90 | 84,097 | 0.05 | 48,420 | 0.20 | 1 | 55 | 44 |
| 1982 | NA | 0.90 | NA | 0.90 | 86,213 | 0.20 | 127,174 | 0.20 | 3 | 50 | 47 |
| 1983 | NA | 0.90 | NA | 0.90 | 134,601 | 0.10 | 124,180 | 0.20 | 5 | 55 | 40 |
| 1984 | NA | 0.90 | NA | 0.90 | 123,190 | 0.10 | 99,592 | 0.20 | 2 | 73 | 25 |
| 1985 | NA | 0.90 | NA | 0.90 | 58,335 | 0.10 | 131,091 | 0.20 | 2 | 38 | 60 |
| 1986 | NA | 0.90 | NA | 0.90 | 23,947 | 0.05 | 168,006 | 0.20 | 2 | 36 | 62 |
| 1987 | NA | 0.90 | NA | 0.90 | 48,972 | 0.10 | 69,900 | 0.20 | 3 | 60 | 37 |
| 1988 | NA | 0.90 | NA | 0.90 | 27,722 | 0.10 | 76,883 | 0.20 | 2 | 44 | 54 |
| 1989 | NA | 0.90 | NA | 0.90 | 141,475 | 0.05 | 156,160 | 0.20 | 1 | 70 | 30 |
| 1990 | NA | 0.90 | NA | 0.90 | 60,230 | 0.05 | 149,377 | 0.20 | 1 | 44 | 55 |
| 1991 | NA | 0.90 | NA | 0.90 | 51,138 | 0.05 | 60,721 | 0.20 | 2 | 46 | 52 |
| 1992 | NA | 0.90 | NA | 0.90 | 95,880 | 0.05 | 113,146 | 0.20 | 1 | 44 | 55 |
| 1993 | NA | 0.90 | NA | 0.90 | 212,757 | 0.05 | 103,531 | 0.20 | 3 | 38 | 59 |
| 1994 | NA | 0.90 | 153,540 | 0.1 | 86,385 | 0.10 | 126,852 | 0.20 | 2 | 57 | 41 |
| 1995 | NA | 0.90 | 184,541 | 0.1 | 61,783 | 0.10 | 68,737 | 0.20 | 4 | 46 | 50 |
| 1996 | NA | 0.90 | 262,852 | 0.1 | NA | 0.05 | 99,677 | 0.20 | 0 | 0 | 0 |
| 1997 | NA | 0.90 | 238,803 | 0.1 | NA | 0.05 | 73,761 | 0.20 | 0 | 0 | 0 |
| 1998 | NA | 0.90 | 211,114 | 0.09 | NA | 0.05 | 112,630 | 0.20 | 0 | 0 | 0 |
| 1999 | NA | 0.90 | 240,002 | 0.05 | 134,048 | 0.10 | 149,410 | 0.20 | 1 | 43 | 55 |
| 2000 | NA | 0.90 | 132,687 | 0.12 | 47,077 | 0.05 | 78,265 | 0.20 | 2 | 13 | 85 |
| 2001 | NA | 0.90 | 105,064 | 0.07 | 53,239 | 0.20 | 60,183 | 0.20 | 3 | 78 | 20 |
| 2002 | NA | 0.90 | 148,465 | 0.17 | 65,611 | 0.10 | 47,332 | 0.20 | 3 | 45 | 52 |
| 2003 | NA | 0.90 | 116,891 | 0.07 | 55,516 | 0.20 | 49,955 | 0.20 | 2 | 29 | 68 |
| 2004 | NA | 0.90 | 118,795 | 0.06 | 83,534 | 0.20 | 51,110 | 0.20 | 5 | 67 | 29 |
| 2005 | NA | 0.90 | 89,072 | 0.10 | 32,098 | 0.20 | 22,852 | 0.20 | 3 | 37 | 59 |
| 2006 | NA | 0.90 | 91,439 | 0.17 | 38,850 | 0.20 | 15,979 | 0.20 | 6 | 59 | 35 |
| 2007 | NA | 0.90 | 59,884 | 0.10 | 27,915 | 0.10 | 14,208 | 0.20 | 5 | 53 | 41 |
| 2008 | 74,919 | 0.10 | 119,808 | 0.11 | NA | 0.05 | 22,156 | 0.20 | 5 | 82 | 13 |
| 2009 | 153,033 | 0.05 | 285,218 | 0.13 | NA | 0.05 | 85,551 | 0.20 | 2 | 64 | 34 |
| 2010 | 61,906 | 0.05 | 72,318 | 0.09 | NA | 0.05 | 48,079 | 0.20 | 3 | 32 | 66 |
| 2011 | 63,628 | 0.05 | 109,335 | 0.08 | NA | 0.05 | 15,599 | 0.20 | 8 | 61 | 30 |
| 2012 | 121,810 | 0.10 | 171,924 | 0.10 | NA | 0.05 | 54,884 | 0.20 | 6 | 75 | 19 |
| 2013 | 116,300 | 0.10 | 224,516 | 0.10 | NA | 0.05 | 75,588 | 0.20 | 6 | 50 | 44 |
| 2014 | 70,470 | 0.05 | 212,201 | 0.12 | NA | 0.05 | 81,502 | 0.20 | 4 | 67 | 29 |
| 2015 | 175,874 | 0.20 | 124,892 | 0.10 | NA | 0.05 | 33,085 | 0.20 | 4 | 30 | 66 |
| 2016 | 88,513 | 0.10 | 96,148 | 0.09 | NA | 0.05 | 35,991 | 0.20 | 4 | 58 | 37 |

Appendix C.– Escapement goals for the Southeast Region, Central Region (Upper Cook Inlet Prince William Sound), Arctic-Yukon-Kuskokwim (AYK) Region, Westward Region (Alaska Peninsula/Aleutian Islands, Kodiak). This table is modified from Munro and Volk 2016. The lower bound (LB) and upper bound (UB) percentages show the change from the LB and UB to the *S*MSY value.

|  |  |  |  |  |  |  |  |  |  |
| --- | --- | --- | --- | --- | --- | --- | --- | --- | --- |
| Area | System | *S*MSY | LB | UB | Type | Initial Year | LB% | UB% | Sources |
| Southeast | Tahltan Lake | 24,000 | 18,000 | 30,000 | BEG | 1993 | 75% | 125% | Humphreys et al. 1994; TTC 1993 |
| Southeast | Redoubt Lake | 17,400 | 10,000 | 25,000 | BEG | 2003 | 57% | 144% | Geiger 2003 |
| Southeast | Chilkoot Lake | 105,000 | 70,000 | 150,000 | BEG | 2009 | 67% | 143% | Eggers et al. 2010 |
| Southeast | East Alsek-Doame River | 16,000 | 13,000 | 26,000 | BEG | 2003 | 81% | 163% | Clark et al. 2003 |
| Southeast | Klukshu River | 9,102 | 7,500 | 11,000 | BEG | 2013 | 82% | 121% | Eggers and Bernard 2011 |
| Southeast | Alsek River | 28,190 | 24,000 | 33,500 | BEG | 2013 | 85% | 119% | Eggers and Bernard 2011 |
| Southeast | Situk River | 50,000 | 30,000 | 70,000 | BEG | 2003 | 60% | 140% | Clark et al. 2002 |
| Upper Cook Inlet | Kasilof River | 240,000 | 160,000 | 340,000 | BEG | 2011 | 67% | 142% | Fair et al. 2010 |
| Upper Cook Inlet | Russian River - Early Run | 36,255 | 22,000 | 42,000 | BEG | 2011 | 61% | 116% | Fair et al. 2010 |
| Prince William Sound | Eshamy Lake | 19,622 | 13,000 | 28,000 | BEG | 2009 | 66% | 143% | Fair et al. 2008 |
| Kuskokwim Area | Middle Fork Goodnews River | 21,890 | 18,000 | 40,000 | BEG | 2007 | 82% | 183% | Brannian et al. 2006; Molyneaux and Brannian 2006 |
| AK Peninsula | Nelson River | 153,000 | 97,000 | 219,000 | BEG | 2004 | 63% | 143% | Nelson et al. 2006 |
| Kodiak | Afognak (Litnik) River | 34000 | 20,000 | 50,000 | BEG | 2005 | 59% | 147% | Nelson et al. 2005 |
| Kodiak | Karluk River Early Run | 175,000 | 110,000 | 250,000 | BEG | 2008 | 63% | 143% | Honnold et al. 2007a |
| Kodiak | Karluk River Late Run | 266,000 | 170,000 | 380,000 | BEG | 2005 | 64% | 143% | Nelson et al. 2005 |
| Kodiak | Upper Station River Early Run | 66,000 | 43,000 | 93,000 | BEG | 2011 | 65% | 141% | Nemeth et al. 2010 |
| Kodiak | Upper Station River Late Run | 186,000 | 120,000 | 265,000 | BEG | 2005 | 65% | 142% | Nelson et al. 2005 |
| Kodiak | Frazer Lake | 118,000 | 75,000 | 170,000 | BEG | 2008 | 64% | 144% | Honnold et al. 2007a |
| Kodiak | Saltery Lake | 23,600 | 15,000 | 35,000 | BEG | 2011 | 64% | 148% | Nemeth et al. 2010 |
| Kodiak | Buskin Lake | 6,650 | 5,000 | 8,000 | BEG | 2011 | 75% | 120% | Nemeth et al. 2010 |
| Average |  |  |  |  |  |  | 68% | 140% |  |